

Research Article

Cite this article: Fischer JH, Bose S, Romero C, Charteris M, Crowe P, Parker GC, Ray S, Rexer-huber K, Sagar PM, Thompson DR, Bell E, Debski I, Quiñones J (2023). Combining tracking with at-sea surveys to improve occurrence and distribution estimates of two threatened seabirds in Peru. *Bird Conservation International*, **33**, e41, 1–12
<https://doi.org/10.1017/S0959270922000442>

Received: 09 May 2022

Revised: 16 August 2022

Accepted: 23 August 2022

Keywords:


Conservation; Geolocation; *Procellaria parkinsoni*; Species distribution models; *Thalassarche salvini*

Author for correspondence:

*Johannes Fischer,

Email: johannesfischer@live.nl

Combining tracking with at-sea surveys to improve occurrence and distribution estimates of two threatened seabirds in Peru

Johannes H. Fischer^{1*} , Samhita Bose¹, Cynthia Romero², Matt Charteris³, Patrick Crowe⁴, Graham C. Parker⁵, Samantha Ray⁴, Kalinka Rexer-huber⁵, Paul M. Sagar⁶, David R. Thompson⁶, Elizabeth Bell⁴, Igor Debski¹ and Javier Quiñones²

¹Aquatic Unit, Department of Conservation, Wellington, New Zealand; ²Oficina de Investigaciones en Depredadores Superiores, Instituto del Mar del Perú, Callao, Perú; ³Waybacks, Charleston, New Zealand; ⁴Wildlife Management International, Blenheim, New Zealand; ⁵Parker Conservation, Dunedin, New Zealand and ⁶National Institute of Water and Atmospheric Research, Wellington, New Zealand

Summary

Seabirds are highly threatened, including by fisheries bycatch. Accurate understanding of offshore distribution of seabirds is crucial to address this threat. Tracking technologies revolutionised insights into seabird distributions but tracking data may contain a variety of biases. We tracked two threatened seabirds (Salvin's Albatross *Thalassarche salvini* $n = 60$ and Black Petrel *Procellaria parkinsoni* $n = 46$) from their breeding colonies in Aotearoa (New Zealand) to their non-breeding grounds in South America, including Peru, while simultaneously completing seven surveys in Peruvian waters. We then used species distribution models to predict occurrence and distribution using either data source alone, and both data sources combined. Results showed seasonal differences between estimates of occurrence and distribution when using data sources independently. Combining data resulted in more balanced insights into occurrence and distributions, and reduced uncertainty. Most notably, both species were predicted to occur in Peruvian waters during all four annual quarters: the northern Humboldt upwelling system for Salvin's Albatross and northern continental shelf waters for Black Petrels. Our results highlighted that relying on a single data source may introduce biases into distribution estimates. Our tracking data might have contained ontological and/or colony-related biases (e.g. only breeding adults from one colony were tracked), while our survey data might have contained spatiotemporal biases (e.g. surveys were limited to waters <200 nm from the coast). We recommend combining data sources wherever possible to refine predictions of species distributions, which ultimately will improve fisheries bycatch management through better spatiotemporal understanding of risks.

Resumen

Las aves marinas están seriamente amenazadas, incluyendo por capturas incidentales en diversas pesquerías. La distribución espacial precisa de aves marinas en zonas oceánicas es crucial para hacer frente a estas amenazas. Las tecnologías de seguimiento satelital revolucionaron la información sobre las distribuciones espaciales de aves marinas, pero estos datos pueden contener diversos sesgos. Rastreamos dos aves marinas amenazadas (Albatros de Salvini *Thalassarche salvini* $n = 60$ y Petrel Negro *Procellaria parkinsoni* $n = 46$) desde sus colonias reproductivas en Aotearoa (Nueva Zelanda) hacia zonas oceánicas de Sudamérica, incluyendo Perú, durante su periodo post reproductivo 2018–2020, de manera simultánea se realizaron siete cruceros científicos de avistamientos de aves marinas en aguas peruanas. Luego se utilizaron variables ambientales y modelos de distribución de especies para predecir su ocurrencia y distribución utilizando una de las fuentes de datos o ambas en combinación. Los resultados muestran diferencias estacionales entre las estimaciones de ocurrencia y distribución cuando se utiliza una sola fuente de datos. Sin embargo, cuando se combinaron ambas fuentes de datos, se obtuvo un resultado mucho más equilibrado con respecto a la ocurrencia y distribución de las especies evaluadas, con una notable disminución del sesgo. En particular, se predijo que ambas especies ocurrirían en aguas peruanas durante todas las estaciones. Donde el Albatros de Salvini se distribuye en Ecosistema de la Corriente de Humboldt, y el Petrel negro en la zona de la plataforma continental al norte del país. Nuestros resultados resaltan que confiar en una sola fuente de datos puede generar un mayor sesgo en las estimaciones de distribución. Nuestros datos de seguimiento satelital podrían tener sesgos ontológicos y/o relacionados al grupo etareo

evaluado en las colonias reproductivas (solo se rastrearon a aves adultas), mientras que nuestros datos de avistamientos a bordo de embarcaciones en Perú, tienen sesgos espaciotemporales (por ejemplo, las evaluaciones se limitaron a aguas <200 nm de la costa). Recomendamos usar ambas fuentes de datos en conjunto, siempre que sea posible, para poder tener una predicción más precisa y fina en la distribución de estas aves marinas, esta información será fundamental para una mejor gestión en el manejo de estas pesquerías para mitigar las capturas incidentales de estas especies a través de una adecuada comprensión de los riesgos a escalas espacio temporales.

Introduction

Seabirds are one of the most threatened species groups on the planet (Dias *et al.* 2019). Seabirds face threats both when breeding on land and while foraging at sea. Marine threats to seabirds include pollution, energy production, mineral exploration and extraction, overfishing, and prominently, bycatch in commercial fisheries. Bycatch is a pervasive and global threat to seabirds. For example, bycatch in longline and trawl fisheries around the world causes the deaths of hundreds of thousands of seabirds, particularly albatrosses and large petrels (e.g. *Procellaria* petrels) (Alfaro-Shigueto *et al.* 2010, Anderson *et al.* 2011). Next to bycatch, overfishing is an ongoing threat to seabirds (Gremillet *et al.* 2018). Mineral exploration and energy production are also threatening seabirds; the latter particularly due to the emerging need for renewable energies (e.g. Warwick-Evans *et al.* 2017). To mitigate these marine threats effectively, robust offshore spatial information during breeding and non-breeding periods is crucial (Carneiro *et al.* 2020).

Tracking technologies revolutionised insights into the offshore distributions of seabirds, and consequently transformed their marine conservation (Burger and Shaffer 2008). Over 28,000 individuals of >200 seabird species have been tracked to date (Bernard *et al.* 2021). Tracking technologies have provided insights into key aspects of seabird biology, including migrations, breeding and non-breeding distributions, marine habitat selection, and offshore behaviour (e.g. Shaffer *et al.* 2006, Rayner *et al.* 2011, Carneiro *et al.* 2020). This information in turn informs the conservation of seabirds by enabling the identification of priority conservation areas (Lascelles *et al.* 2016, Davies *et al.* 2021), the quantification of overlap with threats (e.g. Warwick-Evans *et al.* 2017, Clay *et al.* 2019), and assessments of fisheries interactions (e.g. Orben *et al.* 2021). However, despite the importance of tracking data to seabird conservation, tracking data may contain biases (Carneiro *et al.* 2020, Bernard *et al.* 2021).

Seabird tracking data are often subject to a variety of biases and accounting for these is crucial to ensure conservation efforts are not misdirected. Globally, taxonomic, ontogenic, and spatial biases are prevalent in tracking data (Bernard *et al.* 2021). Heavier seabirds are more frequently tracked than lighter seabirds and coverage is most complete in northern Europe and least complete in tropical latitudes (Mott and Clarke 2018). Seabird tracking is also regularly biased towards certain life-cycle stages and/or colonies. Often, only breeding adults are tracked, neglecting distributions of other life-cycle stages, such as juveniles or non-breeding adults (Carneiro *et al.* 2020). Seabirds can also exhibit inter-colony segregation both at breeding and non-breeding distributions (Gutowsky *et al.* 2015, Bolton *et al.* 2018). Consequently, tracking seabirds from a single colony, often the most accessible one, can create additional biases. Fundamentally, tracking data can be biased due to the selection of individuals that are tracked and the timeframe these data cover (Watanuki *et al.* 2016).

Aside from tracking, traditional vessel-based at-sea surveys or aerial surveys can also provide insights into the offshore distributions of seabirds and inform their marine conservation, but these

methods are also subject to a variety of biases (e.g. Oppel *et al.* 2012, Mott and Clarke 2018). Unlike tracking studies, at-sea or aerial surveys are not affected by ontogenic or colony-related biases, as direct observations allow the recording of all individuals present in an area. Yet, surveys can have spatial or temporal biases as well, which often relate to human resource constraints. For example, surveys provide data within pre-defined areas only. Additionally, surveys may be subject to identification errors, which may be exacerbated in challenging weather conditions (Schaefer *et al.* 2015). Fundamentally, survey data can be biased due to the selection of the area surveyed, the timeframe these data cover, and the taxonomic resolution of this data (Watanuki *et al.* 2016).

Given the variety of shortcomings in either methodology, combining tracking studies with surveys can allow more complete assessments of seabird distributions (Louzao *et al.* 2009, Priddel *et al.* 2014, Sansom *et al.* 2018, Carroll *et al.* 2019). Specifically, combined approaches have provided more robust predictions of distribution (e.g. Priddel *et al.* 2014), estimates of bycatch risk (e.g. Zydelski *et al.* 2011), and identification of marine hotspots (e.g. Yamamoto *et al.* 2015, Carroll *et al.* 2019). However, such dual approaches are uncommon and have largely focused on breeding distribution comparisons within temperate regions (Louzao *et al.* 2009, Sansom *et al.* 2018, Carroll *et al.* 2019), even though other regions are of equal or higher conservation interest. Furthermore, dual approaches usually provide side-by-side comparisons of the two data sources (e.g. Priddel *et al.* 2014, Sansom *et al.* 2018), rather than fully integrating these, for example through species distribution models (SDMs) (Yamamoto *et al.* 2015, Watanuki *et al.* 2016).

Marine conservation of seabirds is increasingly aided by using distributional data in combination with SDMs (Oppel *et al.* 2012). These versatile models can be fitted to tracking data and/or observational data from surveys together with environmental data to predict the occurrence (when distributional data are provided in presence-absence format) or abundance (when data are provided in absolute numbers) of seabirds as a function of the species' ecological niche (e.g. Kruger *et al.* 2018, Cleasby *et al.* 2020, Raine *et al.* 2021). SDMs can provide inference of distribution, occurrence, and abundance, the identification of areas of conservation priority, and threat assessments over large spatial and temporal scales, even though tracking and survey data are usually not spatially and temporally comprehensive (e.g. Watanuki *et al.* 2016). However, the utility of SDMs is dependent on the quality of the distributional and environmental data provided (e.g. Derville *et al.* 2018, Goetz *et al.* 2022). As such, SDM outputs are vulnerable to the various sources of biases inherent to tracking and/or surveys data and benefit from integrated approaches.

Here, we use tracking, at-sea surveys, and SDMs to infer non-breeding occurrence and distribution of two threatened seabirds in Peruvian waters. Our study focused on two species that are listed on the Agreement on the Conservation of Albatrosses and Petrels (ACAP 2009a,b), considered "Vulnerable" on the IUCN Red List (Birdlife International 2021a,b), and are highly susceptible to

bycatch (Richard *et al.* 2020): Salvin's Albatross (*Thalassarche salvini*) and Black Petrel (*Procellaria parkinsoni*). Accurately quantifying their occurrence and distribution, including within Peruvian waters, is of high conservation concern. We therefore tracked both species from Aotearoa (New Zealand) to Peru in 2018–2020, while simultaneously surveying in Peruvian waters. We then sourced a range of environmental variables to model species occurrence and distribution within a Bayesian framework using each data source separately, and both data sources combined. Our aim was to compare estimates and improve predictions of occurrence and distributions by combining data sources to, ultimately, ensure that conservation efforts, such as fisheries bycatch reduction initiatives, are directed as accurately and efficiently as possible.

Methods

Study species and study area

Both Salvin's Albatross and Black Petrel breed in Aotearoa and migrate across the Pacific Ocean to spend their non-breeding period in waters off the west coast of South America, including Peruvian waters. Salvin's Albatrosses are essentially endemic to southern Aotearoa (>99% of the world population) and breed on remote subantarctic islands: Hauriri (Bounty Islands: ~27,000 breeding pairs) (Taylor 2000, Baker and Jenz 2019) and Tine Heke (Snare Islands, Western Chain: ~1,500 pairs) (Baker *et al.* 2015). Black Petrels are endemic to northern Aotearoa and breed on Aotea (Great Barrier Island: ~4,500 pairs) and Hauturu-o-Toi (Little Barrier Island: ~620 pairs) (Bell *et al.* 2016a,b). Salvin's Albatrosses breed from August to April, while Black Petrels breed from October to May (Zhang *et al.* 2020; Rexer-Huber *et al.* 2021). After breeding, Salvin's Albatrosses migrate via a southerly route, exploiting prevailing westerlies, towards Chile, with some birds moving northwards to occupy Peruvian waters during the non-breeding period (Figure 1A) (Thompson *et al.* 2014). Black Petrels also exploit westerlies, but migrate further north towards Galápagos, ultimately spending their non-breeding period in northern South American waters, including in Peruvian waters (Figure 1B) (Bell *et al.* 2020). Both species are highly vulnerable to bycatch in trawl, pelagic longline, and demersal longline fisheries, including in Peruvian waters (Richard *et al.* 2020, Moreno and Quiñones 2022).

Peruvian waters span ~3,000 km from -3.38° S in the north at the border with Ecuador to -19° S in the Hague triangle in the south. Most Peruvian waters (-6° to -19° S) are dominated by comparatively cold waters, which flow northward and are characterised by wind-induced upwelling (Arntz *et al.* 2006, Bakun and Weeks 2008). This area supports one of the world's largest single-species fisheries (Chavez *et al.* 2008). Further north (-4° to -6° S), a warmer ecotone or transitional area has been proposed, however its limits are not well defined (Ibanez-Erquiaga *et al.* 2018). North of -4° S, the tropical Panamian province with very warm waters dominates. Seabird communities within these Peruvian waters are divided between coastal neritic waters and pelagic oceanic waters. In coastal neritic waters (30–40 nm from shore), communities are dominated by guano birds, but approximately 50 nm from shore, communities are dominated by Procellariiformes, including Salvin's Albatrosses and Black Petrels. We defined our exact study area as waters within the Peruvian Economic Exclusive Zone (EEZ) up to 150 nm offshore (278 km) (Figure 1), effectively covering areas of interest for Procellariiformes following Spear *et al.* (2003).

Seabird tracking

We tracked breeding adult Salvin's Albatrosses ($n = 60$) and Black Petrels ($n = 46$) from their breeding colonies in Aotearoa to their non-breeding ranges in South America in 2018–2020 using a variety of tracking technologies (Sagar *et al.* 2018, Bell *et al.* 2020, Thompson *et al.* 2020). Specifically, we equipped Salvin's Albatrosses with 18 Global Positioning System (GPS) tags, 12 Platform Transmitting Terminal (PTT) tags, and 54 Global Location Sensing (GLS) tags between 2018 and 2019 at their breeding colony on Proclamation Island on Hauriri (-47.749° S, 179.028° E) (Table 1). We equipped Black Petrels with 55 GLS tags in 2018 at their breeding colony on Aotea (36.181° S, 175.413° E). We attached GPS and PTT tags on back feathers using pre-cut baseplates, cable ties, Tesa tape, and epoxy glue, while we attached GLS tags on leg-bands using cable ties (Sagar *et al.* 2018, Bell *et al.* 2020). All tags, including attachment apparatuses, were well below the commonly accepted 3% bodyweight threshold for seabird tracking studies (Phillips *et al.* 2004), and we did not observe any negative impacts from the tags deployed (Sagar *et al.* 2018, Bell *et al.* 2020).

The different tracking technologies provided us with different types of data. The GPS and PTT tags on Salvin's Albatrosses were transmitting tags (see https://docnewzealand.shinyapps.io/albatross_tracker/), which were programmed differently depending on the model and make used (recording regimes ranged from two locations per day for PTT tags to ~20 locations per day for some GPS tags). The error for locations obtained from GPS tags was 30–170 m (Irvine *et al.* 2020), while the error for locations obtained from PTT tags was 50–20,000 m (Hazel 2009), depending on the location class. One GPS tag failed almost immediately after deployment and we obtained 17 GPS tracks and 12 PTT tracks (Table 1; Figure 1). However, GPS and PTT tags stopped transmitting partway through the non-breeding period, when birds moulted or when batteries ran out, and thus, these tags did not yield year-round datasets. GLS tags are non-transmitting, so birds must be recaptured to acquire data, but once recaptured, these tags provided year-round datasets. GLS tags were programmed to record light lux every 1–10 minutes, saltwater immersion on a constructed scale every 10–20 minutes, and sea surface temperature (SST) (in °C) when immersed in saltwater for >20 minutes, while saving values every eight hours). We recaptured Salvin's Albatrosses and Black Petrels in 2019 at their breeding colonies and collected 33 and 46 GLS tags, respectively, which yielded 31 and 46 GLS datasets, respectively, as two tags failed (Table 1; Figure 1).

To infer locations from data recorded by GLS tags, we applied the threshold method, followed by an iterative forward step selection with a twilight model, a movement model, and spatial masks using the package “ProbGLS” in R (Merkel *et al.* 2016, R Core Team 2021). We selected a light threshold of 10 for twilight events (Bell *et al.* 2020) and a solar angle window of -7° to -1° for the twilight model (Fischer *et al.* 2021). We applied movement models for dry periods (mean \pm SD = 12 \pm 6 m/s, max = 45 m/s for Salvin's Albatross; 10 \pm 5 m/s, max = 50 m/s for Black Petrel) (Freeman *et al.* 2010, Merkel *et al.* 2016), and wet periods (mean \pm SD = 1 \pm 1.3 m/s, max = 5 m/s for both species) (Merkel *et al.* 2016). We also applied sea-ice, land, and SST spatial masks. For the latter, we cross-referenced SST values recorded by GLS tags with satellite-recorded SST values (Reynolds *et al.* 2007) and allowed the satellite-derived values to differentiate 3° C from GLS records. We then estimated median geographical tracks by generating a cloud of possible locations (1,000 locations per step), selecting the most probable location, and repeating this process for 100 iterations. This

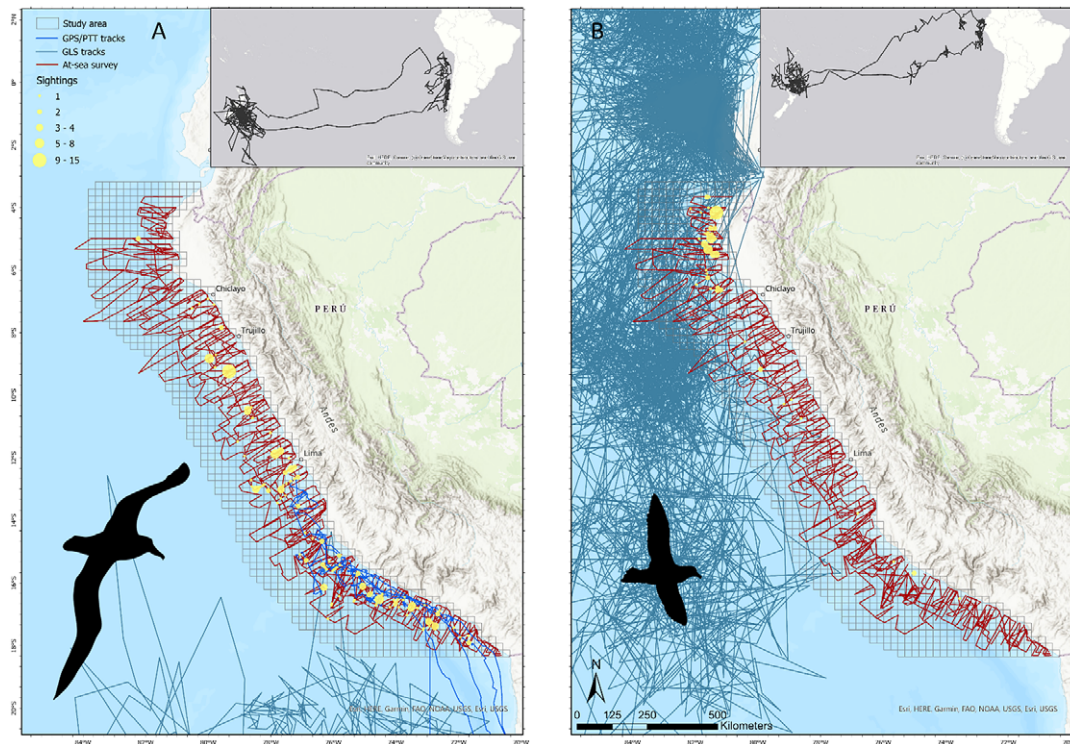


Figure 1. GPS, PTT, and GLS tracks in relation to at-sea survey effort in Peruvian waters for Salvin's Albatrosses (A) and Black Petrels (B). Insets show an example of a year-round GLS track for each species, illustrating the annual migrations from Aotearoa to South America.

Table 1. Summary of tracking effort of Salvin's Albatrosses from Hauriri (Bounty Islands) and Black Petrels from Aotea (Great Barrier Island), Aotearoa (New Zealand). GLS = Global Location Sensing, GPS = Global Positioning System, PTT = Platform Transmitting Terminal.

Species	Age	Tagging location	Year	Tracking technology	Model, make, manufacturer address	Weight (g; % body weight)	<i>n</i> deployed	<i>n</i> retrieved (%)	<i>n</i> datasets obtained (%)
Salvin's Albatross	Adult	Hauriri, Aotearoa	2018/19	GLS	Intigeo-C330, Migrate Technology, Cambridge, UK.	3.3 (0.09%)	32	20 (63%)*	18 (90%)
	Adult	Hauriri, Aotearoa	2018/19	GLS	MK3006, Biotrack, Cambridge, UK.	2.5 (0.07%)	22	13 (59%)*	13 (100%)
	Adult	Hauriri, Aotearoa	2018/19	GPS	Rainier-S20, Wildlife Computers, Redmond, WA, USA.	20 (0.57%)	7	-	7 (100%)
	Adult	Hauriri, Aotearoa	2018/19	GPS	PinPoint, Lotek, Havelock North, Aotearoa, New Zealand.	40 (1.14%)	7	-	6 (86%)
	Adult	Hauriri, Aotearoa	2019/20	GPS	GT-12GS-GPS, Geotrak, Apex, NC, USA.	22 (0.63%)	4	-	4 (100%)
	Adult	Hauriri, Aotearoa	2019/20	PTT	TAV-2630, Telonics, Mesa, AZ, USA.	35 (1.00%)	12	-	12 (100%)
Black Petrel	Adult	Aotea, Aotearoa	2018/19	GLS	Intigeo-C330, Migrate Technology, Cambridge, UK.	3.3 (0.47%)	55	46 (84%)*	46 (100%)

*The lower retrieval rate of GLS tags for Salvin's Albatrosses is likely an artefact of recapture effort as Aotea is much more accessible than Hauriri.

approach allowed us to estimate locations of birds year-round, including during equinoxes (with an approximate error of 145 km) (Merkel *et al.* 2016).

We used the locations obtained from GPS, PTT, and GLS tags to generate presence-absence data for our SDMs (e.g. Goetz *et al.*

2022) within our study area. We first created a minimum convex polygon (MCP) per species around all obtained locations with a 1,000 km buffer. Within the MCPs, we randomly sampled 10 pseudo-absences per obtained location, while accounting for land avoidance. We then resampled the resulting presence-absence

Table 2. Summary of at-sea survey effort in Peruvian waters.

Year	Time period	Days at sea	Latitudinal range (min; max)	Max. distance offshore nm (km)
2018	March	25	-17.62; -5.00	195 (362)
	October–November	45	-18.32; -4.00	138 (255)
2019	February–March	35	-17.79; -3.67	93 (174)
	May–June	33	-18.20; -6.61	99 (184)
	October–November	40	-18.26; -4.00	112 (208)
2020	February–March	34	-17.98; -3.67	99 (184)
	October–November	33	-18.22; -7.14	121 (224)

data per annual quarter (i.e. January–March, April–June, July–September, and October–December) at a 25 × 25 km resolution. We chose this resolution as a compromise between high-resolution GPS, PTT, and at-sea survey (see below) data and lower resolution GLS data.

At-sea surveys

We completed seven semi-standardised vessel-based seabird surveys in 2018–2020 in our study area (Table 2; Figure 1). Our surveys consisted of ~50 uniformly parallel survey transects, separated by 15 nm (28 km), which included coastal-neritic waters and oceanic-pelagic waters up to 120 nm (222 km) from shore, effectively covering the entirety of the Peruvian coast (-18.3° S to -3.7° S) and areas of interest following Spear *et al.* (2003). Effort between surveys was highly comparable. Surveys took 35 days on average and were conducted in three out of four annual quarters (no surveys were conducted during July–September). During these surveys, all seabirds and their GPS locations in 90° quadrants were recorded during continuous strip transects at seven-minute intervals at 14–16 m height at a cruising speed of 10 knots (equating to transect lengths of 1 nm) from dawn till dusk. Seven well-trained observers participated in our at-sea surveys. To overcome identification challenges, albatrosses and petrels were photographed wherever possible. Using photographs, we aged Salvin’s Albatrosses (sub-adults/adults), but Black Petrels cannot be aged phenotypically. Records that were not identifiable to species level were not included in our analyses (on average, 8.6% and 26.6% of records per survey for *Thalassarche* Albatrosses and *Procellaria* Petrels, respectively). We then used records from these surveys to generate presence-absence data for three out of four annual quarters for Salvin’s Albatrosses and Black Petrels (regardless of age). We resampled these data at a 25 × 25 km resolution, matching our tracking data, allowing for direct comparison of both datasets.

Environmental variables

We sourced static and dynamic environmental variables for our SDMs, following previous studies (e.g. Oppel *et al.* 2012, Kruger *et al.* 2018, Degenford *et al.* 2021). Specifically, we sourced bathymetry from the General Bathymetric Chart of the Oceans (GEBCO); <https://www.gebco.net/>), distance to inner shelf (i.e. the inner boundary of the continental shelf, starting at the shore face) and outer shelf (i.e. the outer boundary of the continental shelf at the start of the shelf break) from [marinerregions.org](https://www.marinerregions.org/downloads.php) (<https://www.marinerregions.org/downloads.php>) and SST, chlorophyll-*a* concentration, salinity, sea-surface height (SSH), northern and eastern current velocity,

and turbidity from the Copernicus Marine Service (<https://resources.marine.copernicus.eu/products>) for Peruvian waters up to 150 nm offshore (see Table S1). We sourced dynamic variables at a monthly resolution for the duration of the study period and then averaged these per annual quarter. All environmental variables were resampled to a 25 × 25 km resolution, matching our presence-absence data. We assessed correlation among environmental variables using Spearman regression coefficients (*r*). We excluded variables that showed high correlation (*r* > 0.7) with other variables: distance to inner shelf, chlorophyll-*a* concentration, and SSH (see Table S2). As a final step, we z-transformed all environmental variables to facilitate direct comparison of effects and ease model fit.

Species distribution models

We constructed Bayesian Generalised Linear Models (GLMs) to predict occurrence and distributions of Salvin’s Albatrosses and Black Petrels within Peruvian waters per annual quarter using tracking data only, at-sea survey data only, and both data sources combined. Specifically, per species, we fitted three binomial GLMs with Bernoulli error terms and logit-link functions (e.g. Oppel *et al.* 2012) to our presence-absence (dependent variables) and environmental data (explanatory variables): a GLM using tracking data only, a GLM using at-sea survey data only, and a GLM using presence-absence data from both sources combined. All GLMs had the following structure:

$$\begin{aligned}
 \text{logit}(P_{ij}) = & \alpha + \theta_{\beta}^Q \cdot Q + \beta_{bat} \cdot bat_i + \beta_{d.out} \cdot d. \\
 & out_i + \beta_{sst} \cdot sst_{ij} + \beta_{sal} \cdot sal_{ij} + \beta_{n.vel} \cdot n.vel_{ij} + \beta_{e.vel} \cdot e. \\
 & vel_{ij} + \beta_{turb} \cdot turb_{ij} + \beta_{turb^2} \cdot turb_{ij}^2,
 \end{aligned}$$

in which P_{ij} refers to the occurrence probability per 25 × 25 km cell *i* in quarter *j*, α is the intercept, θ_{β}^Q is a vector of coefficients for the seasonal effect per quarter *j*, Q is a design matrix of the quarterly factor ranging from quarter 2 (April–June) to quarter 4 (October–December), with quarter 1 (January–March) as the reference level, β_{bat} and $\beta_{d.out}$ are the effects of bathymetry and distance to outer shelf, respectively, bat_i and $d.out_i$ are the bathymetry and distance to outer shelf per cell *i*, β_{sst} , β_{sal} , $\beta_{n.vel}$, $\beta_{e.vel}$, and β_{turb} are, respectively, the effects of SST, salinity, northern and eastern current velocity, and turbidity, sst_{ij} , sal_{ij} , $n.vel_{ij}$, $e.vel_{ij}$, $turb_{ij}$ are, respectively, the SST, salinity, northern and eastern current velocity, and turbidity per cell *i* in quarter *j*, β_{turb^2} is the quadric effect of turbidity, and $turb_{ij}^2$ is the quadratic of turbidity. We explored other quadratic effects (e.g. for SST), but these led to poor model convergence and therefore only one quadratic was included. Due to small sample

size constraints, we did not incorporate additional random effects (e.g. for individual or tag type). We used these GLMs to estimate the quarterly occurrence probability per cell per species as well as the influence of environmental variables on occurrence. We used GLMs instead of more advanced generalised additive models, machine-learning techniques, or ensemble models (e.g. Oppel *et al.* 2012, Kruger *et al.* 2018) to facilitate the direct and straightforward comparison of results, including linear coefficients, based on tracking data only, survey data only, and both data sources combined.

We fitted our GLMs using Markov chain Monte Carlo (MCMC) algorithms within the Bayesian modelling program OpenBUGS (Spiegelhalter *et al.* 2007), allowing all sources of error to be propagated into posterior distributions. We used vague priors for intercepts ($\alpha = N[0, 0.001]$) and coefficients for static and dynamic variables ($\beta = N[0, 0.01]$), but we somewhat restricted priors for coefficients of quarterly effects ($\beta = N[0, 0.1]$) as we did not have at-sea survey data for each quarter. We pooled two MCMC chains with 20,000 iterations, after a burn-in of 20,000 iterations, which was sufficient to reach convergence, as assessed through the Gelman–Rubin statistic ($\hat{R} < 1.05$) and visual inspection of trace plots. We report the means of posterior distributions with 95% credible intervals (CIs).

Results

Seabird tracking

In 2018–2020, 17 GPS, 12 PTT, and 31 GLS tracks were obtained from Salvin's Albatrosses tagged at Hauriri (Table 1). Of these tracked birds, 29%, 0%, and 6%, respectively, spent time in Peruvian waters (Figure 1A). Tracked Salvin's Albatrosses were present in Peruvian waters from February to July. In addition, 46 Black Petrel tracks were obtained, of which 63% spent time in Peruvian waters (Figure 1B). Tracked Black Petrels were present in Peruvian waters from April to October. These tracks translated into recorded

presence in 125 and 211 25×25 km cells for Salvin's Albatross and Black Petrel, respectively, used in SDMs.

At-sea surveys

The at-sea surveys between 2018 and 2020 recorded a total of 292 Salvin's Albatrosses and 57 Black Petrels in Peruvian waters (Figure 1). Salvin's Albatross records ranged from February to November. Ontogenic ratios for Salvin's Albatrosses differed considerably throughout the year. During February–March, adults ($n = 60$; 86%) were more prevalent than subadults ($n = 10$; 14%). However, during October–November, subadults ($n = 18$; 62%) were more prevalent than adults ($n = 11$; 38%). Black Petrel records ranged from February to October. Data from these at-sea surveys resulted in recorded presence in 63 and 33 25×25 km cells for Salvin's Albatross and Black Petrel, respectively, used in SDMs.

Predictions of occurrence and distribution

Predictions of occurrence and distribution varied depending on the data used in our SDMs (Figures 2–4). For Salvin's Albatross, when considering tracking data only, the mean occurrence probability per 25×25 km cell was estimated at 0.09 (CI = 0.05–0.14), with a peak value in April–June and a low value in October–December (Figure 2A). Tracking data predicted relatively high Salvin's Albatross occurrence in southern Peruvian waters around the continental shelf, the continental slope, the Peru–Chile trench within the Humboldt upwelling system, and along the northern Nazca Ridge (approximately -11° to -18° S) as well as in a small hotspot in northern Peruvian waters over the continental shelf (-4° to -8° S) (Figure 3A–D). When considering at-sea survey data only, mean occurrence probability per cell was estimated at 0.11 (0.04–0.33), with a peak value in April–June remaining comparatively high in October–December, and a low value in January–March (Figure 2A). The high uncertainty surrounding the July–September estimate was caused by the absence of surveys during this quarter. At-sea survey

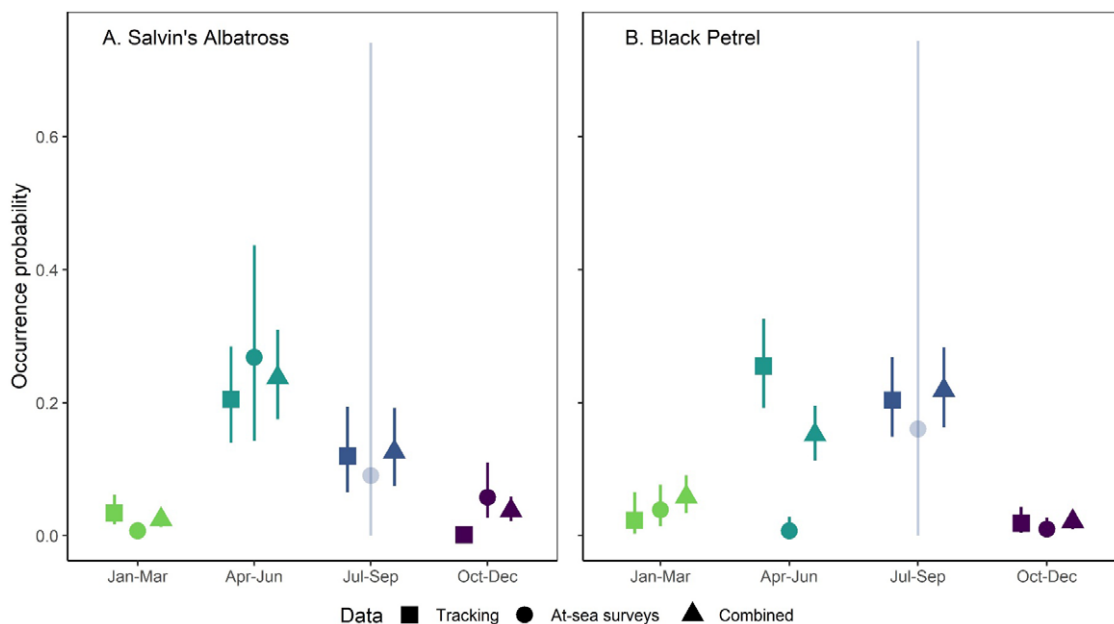


Figure 2. Estimates of quarterly occurrence probability per 25×25 km cell for Salvin's Albatross (A) and Black Petrel (B) in Peruvian waters, based on different data sources. Note no at-sea surveys were conducted in July–September (translucent symbols).

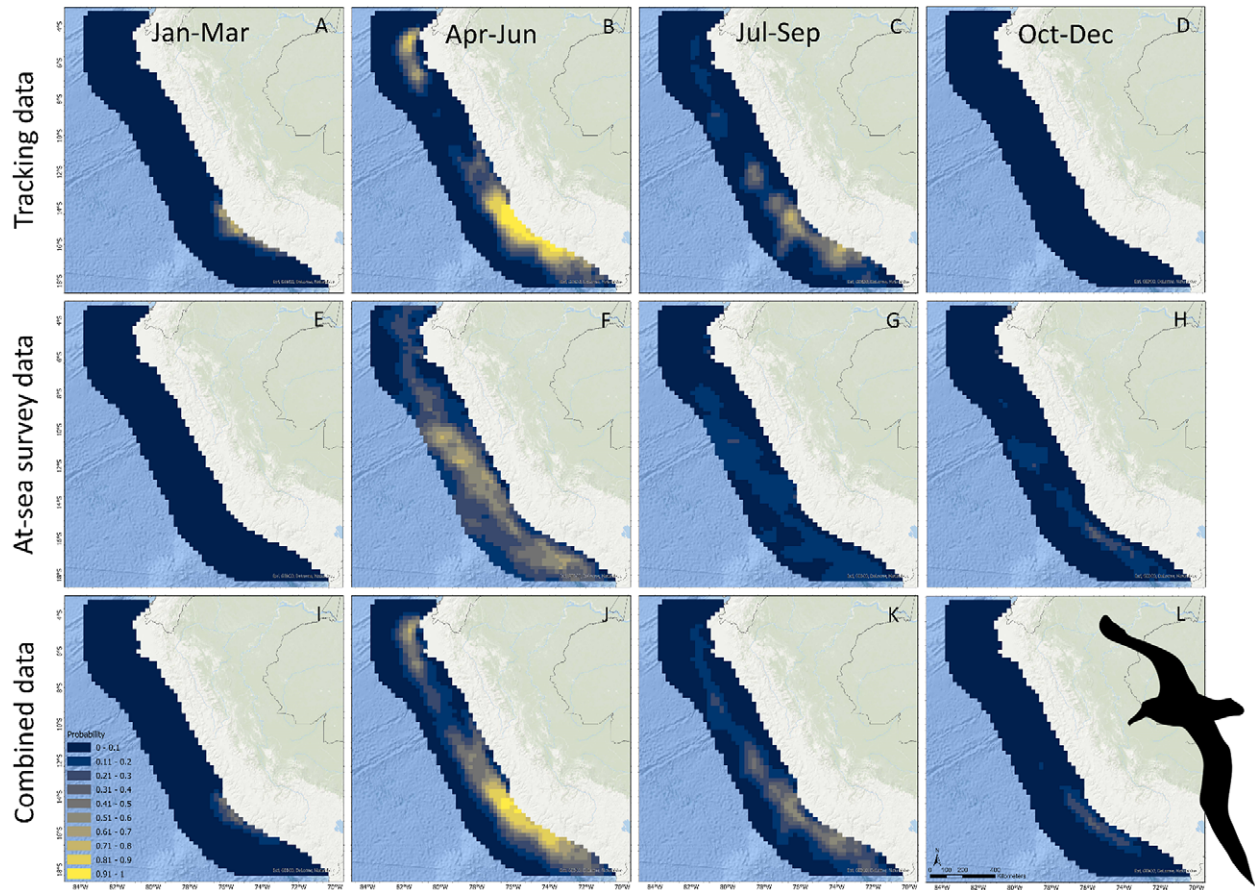


Figure 3. Predictions of Salvin's Albatross occurrence in Peruvian waters based on tracking data (A–D), at-sea survey data (E–H), and both data sources combined (I–L). Note no at-sea surveys were conducted during July–September.

data predicted relatively high Salvin's Albatross occurrence within the Humboldt upwelling system and central Peruvian waters (-14° to -17° S), but no occurrence in northern Peru (Figure 3E–H). Differences between seasonal effects depending on the data used are further highlighted in contrasting seasonal β estimates (Figure 5A). Combining the two data-sources resulted in a balance between the two independent estimates and reduced uncertainty in the estimated mean occurrence probability per cell (0.11; 0.07–0.15). Occurrence reached a peak in April–June and was lowest in October–December through to January–March (Figure 2A). Combined data predicted Salvin's Albatross occurrence in southern Peru in the Humboldt upwelling system, the continental slope, the Peru–Chile trench, and the abyssal plain waters beyond (-12° to -18° S) during all four annual quarters, as well as seasonal use of waters around the continental shelf further north (-4° to -8° S) (Figure 3I–L).

For Black Petrels, when considering tracking data only, mean occurrence probability per 25×25 km cell was estimated at 0.13 (CI = 0.09–0.18), with peak values estimated in April–June and July–September and the lowest value estimated in January–March (Figure 2B). Tracking data predicted relatively high Black Petrel occurrence offshore in the Peru Basin (approximately -4° to -14° S) and inshore waters in northern Peru (-4° to -5° S) (Figure 4A–D). When considering at-sea survey data only, however, mean estimated occurrence probability per cell was low (0.05; 0.004–0.22) and more constant, but occurrence showed a peak value in January–March (Figure 2B). The high uncertainty in July–September was caused by the absence of surveys during this quarter. At-sea survey

data predicted relatively high Black Petrel occurrence in northern Peruvian waters along the continental shelf (-4° to -6° S), but notably not offshore within the Peru Basin (Figure 4E–H). Differences between seasonal effects depending on the data used are further highlighted in contrasting seasonal β estimates (Figure 5B). Combining both data sources resulted in a balance between the two independent estimates and reduced uncertainty in the estimated mean occurrence probability per cell (0.11; 0.08–0.15). Black Petrel occurrence reached a peak in April–June and in July–September and was lowest in October–December and in January–March. Combining data sources predicted localised Black Petrel occurrence in northern Peruvian inshore waters over the continental shelf during all four annual quarters (-4° to -5° S), as well as seasonal use of offshore waters in the Peru Basin (-5° to -10° S) (Figure 4I–L).

Influence of environmental variables

Environmental variables influenced occurrence predictions differently depending on the data used in our SDMs (Figure 5). For Salvin's Albatross, tracking data showed that the probability of occurrence was negatively influenced by northern current velocity, SST, and distance to outer shelf, i.e. Salvin's Albatrosses in Peru favoured cold, slow flowing waters close to the outer shelf. These results contrast with estimates based on at-sea survey data, which showed that probability of occurrence had a positive relationship with eastern current velocity and a quadratic relationship with turbidity. Thus, at-sea survey data suggested that Salvin's

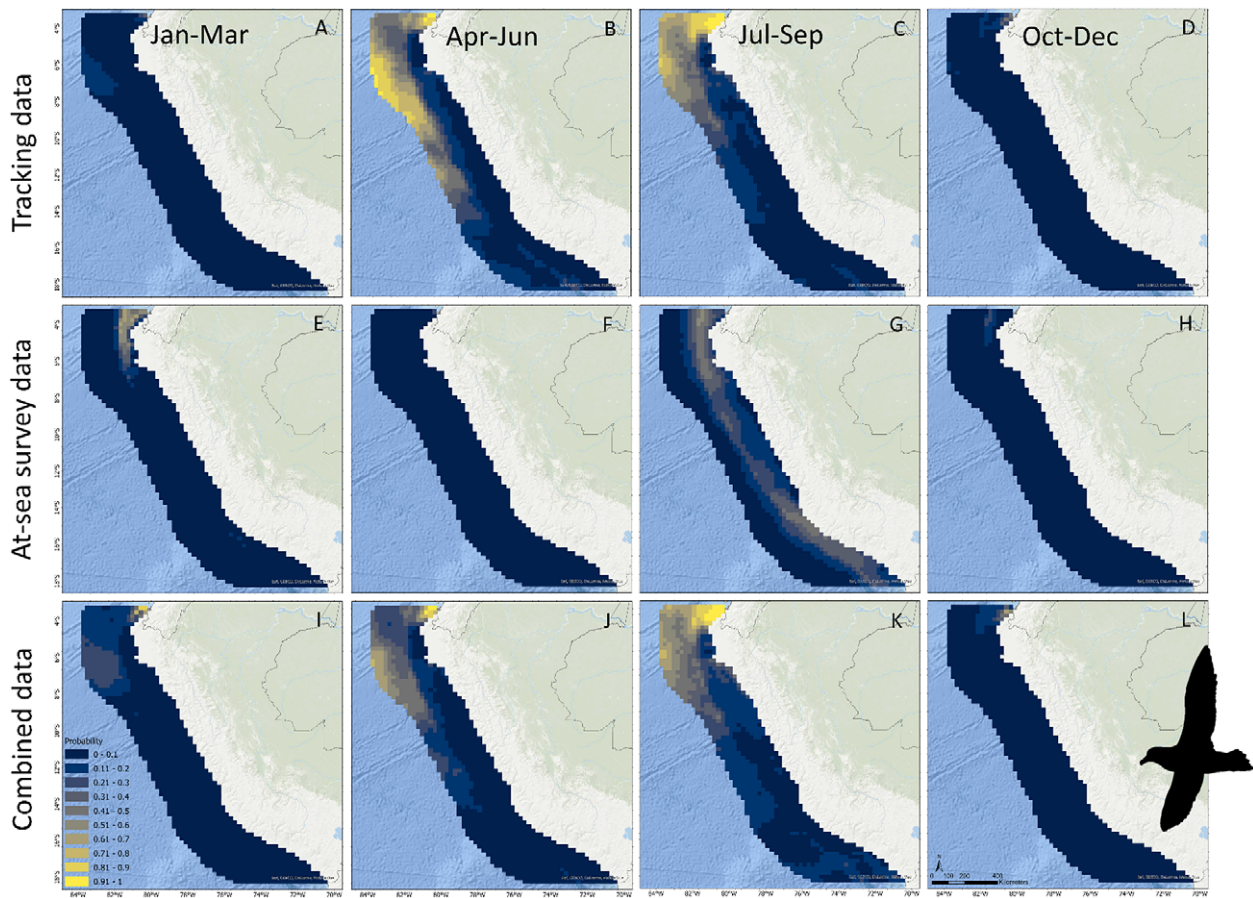


Figure 4. Predictions of Black Petrel occurrence in Peruvian waters based on tracking data (A–D), at-sea survey data (E–H), and both data sources combined (I–L). Note no at-sea surveys were conducted in July–September.

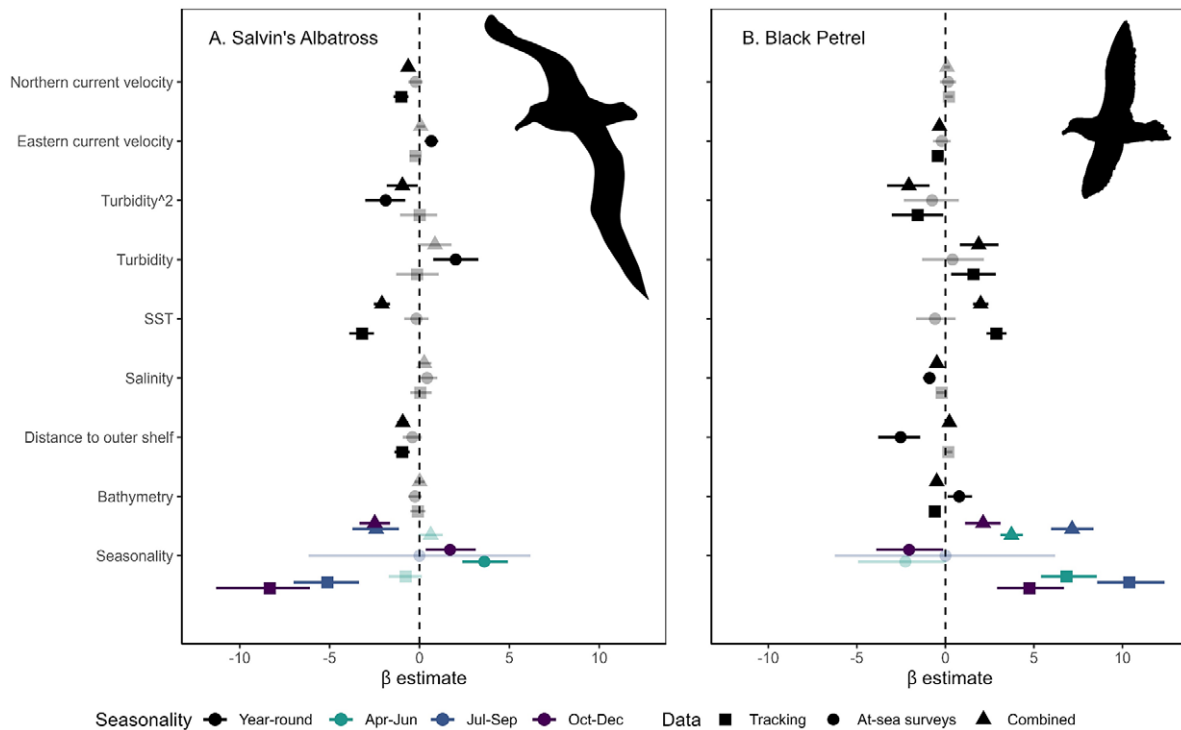


Figure 5. Estimates of coefficients used to estimate occurrence of Salvin's Albatrosses (A) and Black Petrels (B) in Peruvian waters, using different presence-absence data sources. All environmental coefficients indicate year-round relationships (black symbols), while seasonal effects are modelled using separate coefficients (coloured symbols). Translucent symbols indicate CIs include zero.

Albatrosses favour faster flowing waters at a turbidity optimum. When modelling Salvin's Albatross occurrence using both data combined, important interactions with environmental variables were negative relationships with northern current velocity, distance to outer shelf, and a potential quadratic relationship with turbidity. Therefore, combined data suggested that Salvin's Albatrosses favoured cold, slow flowing waters close to the outer shelf, potentially at a turbidity optimum.

Tracking data for Black Petrels showed that the probability of occurrence had a negative relationship with eastern current velocity and bathymetry, a positive relationship with SST, and a quadratic relationship with turbidity. Tracking data thus suggested that Black Petrels in Peru favoured deep, warm, slow flowing waters, at a turbidity optimum. At-sea survey data showed that the probability of Black Petrel occurrence had a negative relationship with salinity and distance to outer shelf and a positive relationship with bathymetry, i.e. Black Petrels favoured shallow waters near the outer shelf with low salinity. When modelling Black Petrel occurrence using both data sources combined, important interactions with environmental variables were negative relationships with eastern current velocity, salinity, and bathymetry, positive relationships with SST and distance to outer shelf, and a quadratic relationship with turbidity. Combined data thus suggested that Black Petrels favoured deep, warm, slow flowing waters, with low salinity, at a turbidity optimum.

Discussion

Using tracking or at-sea survey data independently in our SDMs showed seasonal differences in estimates of species occurrence and distribution and contrasting interactions with environmental variables, while combining both data sources resolved spatiotemporal imperfections, provided more balanced estimates, and reduced uncertainty. Inferences from tracking data showed that Salvin's Albatrosses and Black Petrels occurred in Peruvian waters during February–July and April–October, respectively, while inferences from at-sea survey data showed that occurrences extended from February to November and from February to October for Salvin's Albatross and Black Petrel, respectively. Tracking data highlighted different hotspots of relatively high occurrence when compared with at-sea survey data (seasonal occurrence in northern Peruvian waters for Salvin's Albatross and seasonal occurrence in the Peru Basin for Black Petrels). Underlying these differences are contrasting estimates of interactions with environmental variables based on the data source used. Combining both data sources resulted in more balanced estimates of occurrence, distribution, and habitat selection. Most notably, combining both data sources predicted occurrence of both species in Peruvian waters during all four annual quarters (in the northern Humboldt upwelling system for Salvin's Albatross and the equatorial and tropical superficial waters in northern Peru over the continental shelf for Black Petrels).

The differences between tracking-based and survey-based results may be caused by ontogenic biases in tracking data. Tracking studies usually focus on breeding adults (e.g. Peron and Gremillet 2013), as was the case in our study. The higher occurrence of both species in Peruvian waters during April–September is a likely consequence of the seasonal migrations of breeding adults from Aotearoa to South America (Spear *et al.* 2003, Quinones *et al.* 2021). However, juveniles and non-breeding adults account for 47–81% of Procellariiform seabird populations (Carneiro *et al.* 2020). Excluding these life-cycle stages from tracking can bias occurrence and

distribution estimates. We observed predominantly subadult Salvin's Albatrosses during October–November, which suggests that life-cycle stages, other than breeding adults, utilise Peruvian waters at different times. Consequently, ontogenic biases in tracking data may have caused our contrasting results. Our study thus highlights the importance of tracking life-cycle stages beyond breeding adults (de Grissac *et al.* 2016, Carneiro *et al.* 2020, Frankish *et al.* 2021). Once other life-cycle stages have been tracked, future research could repeat our efforts and provide separate estimates for both adults and subadults. Additionally, as our estimates were only quarterly, future research should quantify occurrence and distribution of these species in Peruvian waters at a higher temporal resolution (e.g. monthly).

The differences between tracking-based and survey-based results could also be explained by geographical biases in both data sources. Seabirds can exhibit inter-colony segregation in non-breeding areas (Gutowsky *et al.* 2015, Bolton *et al.* 2018). Yet, tracking studies often track birds from a single breeding colony (i.e. the largest or the most accessible colony), as was the case in our study. Single colony tracking studies can bias occurrence and distribution estimates (e.g. Rayner *et al.* 2011). There is evidence suggesting that Salvin's Albatrosses exhibit inter-colony segregation at their non-breeding distribution. Birds from the relatively small Tine Heke colony may potentially migrate further north than birds from the larger Hauriri colony (Thompson *et al.* 2014). However, it should be noted that this evidence is based on GLS tracking only. The contrasts between GLS and PTT/GPS tracks in our study (Figure 1) suggested that Salvin's Albatross GLS tracks may be subject to local idiosyncrasies in accuracy (Halpin *et al.* 2021), despite our best efforts to correct for these (Merkel *et al.* 2016), and these idiosyncrasies should be further investigated. No evidence for inter-colony segregation exists for Black Petrels (Bell *et al.* 2014). Our results thus also support the need for multi-colony tracking studies. Inversely, while our effort was highly comparable between surveys, effort was concentrated on waters within 120 nm (222 km) from shore, due to logistical constraints. This limitation could have also biased survey results to inshore waters, which, for example, could explain the lack of estimated occurrence of Black Petrels in the Peru Basin. Our results thus also highlight the need for ocean-wide at-sea surveys (Priddel *et al.* 2014).

Peruvian waters form an important part of the distribution of these two vulnerable seabirds. Both species migrate >7,000 km from their breeding colonies in Aotearoa to Peruvian waters and utilise these waters during different life-cycle stages. It is likely that the high productivity in Peruvian waters attracts these birds, and the relationships of both species with turbidity, current velocity, and distance to shelf supported this. The cold waters of the Northern Humboldt Current that dominate Peruvian waters are highly productive. In fact, this area is the most productive eastern boundary current system worldwide (Paulik 1981, Spear *et al.* 2003, Chavez *et al.* 2008). Additionally, eddy-like structures occur in southern Peruvian offshore waters (from -15° S to -17° S), extending the high primary productivity to offshore waters (Chaigneau *et al.* 2008, 2013). The high productivity results in heightened zooplankton concentrations and increased abundances of prey species such as squid (e.g. *Abraliopsis* sp., *Argonauta nouryi*, and *Dosidicus gigas*) and fish (e.g. *Trachurus* sp., *Vinciguerra* sp.) (Acha *et al.* 2015, Quinones *et al.* 2020, 2021). Consequently, the highly productive waters off Peru sustain these two seabirds from Aotearoa. However, these highly productive Peruvian waters are also characterised by high inter-annual climatic fluctuations (e.g. Chavez *et al.* 2008).

Here, we used only three years of tracking and survey data, which were not sufficient to account for inter-annual variation. Therefore, the continuation of tracking and simultaneous at-sea surveys would improve insights into inter-annual variations.

Robust, year-round insights into the distribution and occurrence of vulnerable seabirds are crucial for their conservation. Salvin's Albatrosses and Black Petrels are highly vulnerable to bycatch in commercial trawl, pelagic longline, and demersal longline fisheries, in their breeding ranges surrounding Aotearoa, mainly due to industrial fisheries (Richard *et al.* 2020), and their non-breeding ranges in South America, due to both industrial and artisanal fisheries (Abraham *et al.* 2019, Moreno and Quiñones 2022). In Aotearoa, estimates of seabird distribution and occurrence are overlaid with fishing effort to assess risks and prioritise conservation efforts (e.g. Richard *et al.* 2020). However, such exercises are rare for the non-breeding ranges of species traversing entire ocean basins (Clay *et al.* 2019). Despite some shortcomings, we here provide estimates of the distribution and occurrence of Salvin's Albatross and Black Petrel in their non-breeding range within Peruvian waters. These estimates will facilitate future overlap analyses with commercial (e.g. Kroodsmá *et al.* 2018), and potentially, artisanal longline fishing effort in Peruvian waters (Majluf *et al.* 2002, Doherty *et al.* 2014). Such overlap analyses will allow non-breeding distribution risk assessments and, ultimately, could improve the efficacy of future conservation actions for both species (e.g. targeted promotion of best practice mitigation measures) (ACAP 2021a,b).

Our results highlight the value of using multiple data sources to overcome imperfections and biases prevalent in single data sources. Combining seabird tracking with at-sea surveys allowed us to overcome potential ontogenic and geographical biases in either data source, resulting in more robust occurrence and distribution estimates. Fusing data sources highlighted the importance of Peruvian waters for Salvin's Albatrosses and Black Petrels during all four annual quarters. These insights will prove invaluable for future non-breeding distribution risk assessments and conservation actions for these species. As such, our study exemplifies how trans-oceanic collaborations can provide more robust guidance for seabird conservation throughout their distribution and annual cycles.

Acknowledgements. MC, GCP, KRH, PMS, and DRT thank Steve Kafka, Jim Dilley, and the entire crew of the *Evohe* for safe passage to and from Hauriri and the staff of the DOC Murihiku office for biosecurity and logistical support. PC, SR, and EB thank Ngāti Rehua Ngāti Wai ki Aotea for their ongoing support and permission to carry out this work, L. Mack, the staff of the DOC Okiwi office, S. Matthews, and C. Matthews for logistical support, and J. Ranstead, A. Clow, L. Clow, the crew of the *Southern Cross*, J. Molloy, J. Gower, J. Walter, D. Lees, G. Hobson, J. Dolton, S. van Leeuwen, and H. Ricardo, for assistance in the field on Aotea. CM and JQ thank C. Moreno, M. Manrique, L. Mayaute, J. Calderon, and S. Falero for assisting with at-sea surveys and the skippers and crew of the *Humboldt* and the *Jose Olaya Balandra* for their safe navigation of Peruvian waters. All fieldwork in Aotearoa was carried out with permission from local iwi (Kaitiaki Rōpū ki Murihiku and Ngāi Tahu for Hauriri, Ngāti Rehua Ngāti Wai ki Aotea for Aotea), the New Zealand Department of Conservation (M1819/04 for Hauriri), and Animal Ethics Committees (Gummer 2013a, b for Hauriri, AEC321 for Aotea). Fieldwork on Hauriri was funded by the Conservation Services Program of the New Zealand Department of Conservation (POP2017-03). Fieldwork on Aotea was funded by the New Zealand Ministry of Primary Industries (PRO2017-01A). At-sea surveys in Peruvian waters were funded by the Instituto del Mar del Perú.

Supplementary Materials. To view supplementary material for this article, please visit <http://doi.org/10.1017/S0959270922000442>.

References

- Abraham, E., Richard, Y., Walker, N., Gibson, W., Ochi, D., Tsuji, S., *et al.* (2019) Assessment of the risk of surface longline fisheries in the Southern Hemisphere to albatrosses and petrels, for 2016. In *Report prepared for the 13th Meeting of the Ecologically Related Species Working Group (ERSWG13) of the Commission for the Conservation of Southern Bluefin Tuna (CCSBT-ERS/1905/17)*.
- ACAP. (2009a) Species Assessments: Black Petrel *Procellaria parkinsoni*. Accessed online 04 November 2021 from <http://www.acap.aq>.
- ACAP. (2009b) Species Assessments: Salvin's Albatross *Thalassarche salvini*. Accessed online 04 November 2021 from <http://www.acap.aq>.
- ACAP. (2021a) ACAP Review of Mitigation Measures and Best Practice Advice for Reducing the Impact of Demersal Longline Fisheries on Seabirds. Accessed online 22 July 2022 from <http://www.acap.aq>.
- ACAP. (2021b) ACAP Review of Mitigation Measures and Best Practice Advice for Reducing the Impact of Pelagic Longline Fisheries on Seabirds. Accessed online 22 July 2022 from <http://www.acap.aq>.
- Acha, E. M., Piola, A., Iribarne, O. and Mianzan, H. (2015) *Ecological processes at marine fronts: oases in the ocean*. Heidelberg, Germany: Springer.
- Alfaro-Shigueto, J., Mangel, J., Pajuelo, M., Dutton, P. H., Seminoff, J. A. and Godley, B. J. (2010) Where small can have a large impact: structure and characterization of small-scale fisheries in Peru. *Fish. Res.* **106**: 8–17.
- Anderson, O. R. J., Small, C. J., Croxall, J. P., Dunn, E. K., Sullivan, R. J., Yates, O., *et al.* (2011) Global seabird bycatch in longline fisheries. *Endanger. Species Res.* **14**: 91–106.
- Arntz, W. E., Gallardo, V. A., Gutierrez, D., Isla, E., Levin, L. A., Mendo, J., *et al.* (2006) El Niño and similar perturbation effects on the benthos of the Humboldt, California, and Benguela Current upwelling ecosystems. *Adv. Geosci.* **6**: 243–265.
- Baker, B. G. and Jenz, K. (2019) 2018 *Aerial survey of Salvin's albatross at the Bounty Islands*. Wellington, NZ: Department of Conservation.
- Baker, B., Jenz, K. and Sagar, P. (2015) *Aerial survey of Salvin's albatross at the Snares, Western Chain*. Wellington, NZ: Department of Conservation.
- Bakun, A. and Weeks, S. J. (2008) The marine ecosystem off Peru: what are the secrets of its fishery productivity and what might its future hold? *Prog. Oceanogr.* **79**: 290–299.
- Bell, E. A., Mischler, C. P., McArthur, N. and Sim, J. L. (2016a) *Black petrel (Procellaria parkinsoni) population study on Hauturu-o-Toi/Little Barrier Island, 2015/16*. Wellington, NZ: Department of Conservation.
- Bell, E. A., Mischler, C., Sim, J. L. and Scofield, P. (2016b) *Population parameters of black petrels (Procellaria parkinsoni) on Great Barrier Island (Aotea Island), 2015/16*. Wellington, NZ: Department of Conservation.
- Bell, E. A., Ray, S. and Crowe, P. (2020) *Population trends, at-sea distribution, and breeding population size of black petrels (Procellaria parkinsoni) – 2018/19 operational report*. Wellington, NZ: Fisheries New Zealand.
- Bell, E. A., Sim, J. L., Abraham, E., Torres, L. and Shaffer, S. (2014) *At-sea distribution of the black petrels (Procellaria parkinsoni) on Great Barrier Island (Aotea Island), 2009/10*. Wellington, NZ: Department of Conservation.
- Bernard, A., Rodrigues, A. S. L., Cazalis, V. and Gremillet, D. (2021) Toward a global strategy for seabird tracking. *Conserv. Lett.* **14**: e12804.
- BirdLife International (2021a) Species Factsheet: *Procellaria parkinsoni*. Accessed online 04 November 2021 from <http://www.birdlife.org>.
- BirdLife International (2021b) Species Factsheet: *Thalassarche salvini*. Accessed online 04 November 2021 from <http://www.birdlife.org>.
- Bolton, M., Conolly, G., Carroll, M., Wakefield, E. D. and Caldow, R. (2018) A review of the occurrence of inter-colony segregation of seabirds foraging areas and the implications for marine environmental impact assessment. *Ibis* **161**: 241–259.
- Burger, A. E. and Shaffer, S. A. (2008) Application of tracking and data-logging technology in research and conservation of seabirds. *Auk* **125**: 253–264.
- Carneiro, A. P. B., Pearmain, E. J., Oppel, S., Clay, T. A., Phillips, R. A., Bonnet-Lebrun, A.-S., *et al.* (2020) A framework for mapping the distribution of seabirds by integrating tracking, demography, and phenology. *J. Appl. Ecol.* **57**: 514–525.
- Carroll, M. J., Wakefield, E. D., Scragg, E. S., Owen, E., Pinder, S., Bolton, M., *et al.* (2019) Matches and mismatches between seabird distributions estimated from at-sea surveys and concurrent individual-level tracking. *Front. Ecol. Evol.* **7**: 333.

- Chaigneau, A., Dominguez, N., Eldin, G., Vasquez, L., Flores, R., Grados, C., et al. (2013). Near-coastal circulation in the Northern Humboldt Current System from shipboard ADCP data. *J. Geophys. Res. Oceans* **118**: 5251–5266.
- Chaigneau, A., Gizolme, A. and Grados, C. (2008) Mesoscale eddies off Peru in altimeter records: identification algorithms and eddy spatio-temporal patterns. *Prog. Oceanogr.* **79**: 106–119.
- Chavez, F. P., Bertrand, A., Guevara-Carrasco, R., Soler, P. and Csirke, J. (2008) The northern Humboldt Current System: brief history, present status and view towards the future. *Prog. Oceanogr.* **79**: 95–105.
- Clay, T. A., Small, C., Tuck, G. N., Pardo, D., Carneiro, A. P. B., Wood, A. G., et al. (2019) A comprehensive large-scale assessment of fisheries bycatch risk to threatened seabird populations. *J. Appl. Ecol.* **56**: 1882–1893.
- Cleasby, I. R., Owen, E., Wilson, L., Wakefield, E. D., O'Connell, P. and Bolton, M. (2020) Identifying important at-sea areas for seabirds using species distribution models and hotspot mapping. *Biol. Conserv.* **241**: 108375.
- Davies, T. E., Carneiro, A. P. B., Tarzia, M., Wakefield, E., Hennicke, J. C., Frederiksen, M., et al. (2021) Multispecies tracking reveals a major seabird hotspot in the North Atlantic. *Conserv. Lett.* **14**: e12824.
- De Grissac, S., Borger, L., Guitteaud, A. and Weimerskirch, H. (2016) Contrasting movement strategies among juvenile albatrosses and petrels. *Sci. Rep.* **6**: 26103.
- Degenford, J. H., Liang, D., Bailey, Hoover, A. L., Zarate, P., Azocar, J., et al. (2021) Using fisheries observation data to develop a predictive species distribution model for endangered sea turtles. *Conserv. Sci. Pract.* **3**: e349.
- Derville, S., Torres, L. G., Iovan, C. and Garrigue, C. (2018) Finding the right fit: comparative cetacean distribution models using multiple data sources and statistical approaches. *Divers. Distrib.* **24**: 1657–1673.
- Dias, M. P., Martin, R., Pearmain, E. J., Burfield, I. J., Small, C., Phillips, R.A., et al. (2019) Threats to seabirds: a global assessment. *Biol. Conserv.* **237**: 525–537.
- Doherty, P. D., Alfaro-Shigueto, J., Hodgson, D. J., Mangel, J. C., Witt, M. J. and Godley, B. J. (2014) Big catch, little sharks: insight into Peruvian small-scale longline fisheries. *Ecol. Evol.* **4**: 2375–2383.
- Fischer, J. H., Debski, I., Spitz, D. B., Taylor, G. A. and Wittmer, H. U. (2021) Year-round offshore distribution, behaviour, and overlap with commercial fisheries of a Critically Endangered small petrel. *Mar. Ecol. Prog. Ser.* **660**: 171–187.
- Frankish, C. K., Cunningham, C., Manica, A., Clay, T. A., Prince, S. and Phillips, R. A. (2021) Tracking juveniles confirms fisheries-bycatch hotspot for an endangered albatross. *Biol. Conserv.* **261**: 109288.
- Freeman, R., Dennis, T., Landers, T., Thompson, D., Bell, E., Walker, M., et al. (2010) Black Petrels (*Procellaria parkinsoni*) patrol the ocean shelf-break: GPS tracking of a Vulnerable Procellariiform seabird. *PLoS One* **5**: e9236.
- Goetz, K., Stephenson, F., Hoskins, A., Bindoff, A. D., Orben, R.A., Sagar, P. M., et al. (2022) Data quality influences the predicted distribution and habitat of four southern-hemisphere Albatross species. *Front. Mar. Sci.* **9**: 782923.
- Gremillet, D., Ponchon, A., Paleczny, M., Palomares, M. D., Karpouzi, V. and Pauly, D. (2018) Persistent worldwide seabird-fishery competition despite seabird community decline. *Curr. Biol.* **28**: 4009–4013.
- Gummer, H. (2013a) *Attaching radio and data-storage tags to birds: feather/skin mounts (tape/ties/glue). Standard operating procedure.* Wellington, NZ: Department of Conservation.
- Gummer, H. (2013b) *Attaching radio and data-storage tags to birds: leg mounts. Standard operating procedure.* Wellington, NZ: Department of Conservation.
- Gutowsky, S. E., Leonard, M. L., Conners, M. G., Shaffer, S. A. and Jonsen, I. D. (2015) Individual-level variation and higher-level interpretation of space use in wide-ranging species: an albatross case study of sampling effects. *Front. Mar. Sci.* **2**: 93.
- Halpin, L. R., Ross, J. D., Ramos, R., Mott, R., Carlile, N., Golding, N., et al. (2021) Double-tagging scores of seabirds reveals that light-level geolocator accuracy is limited by species idiosyncrasies and equatorial solar profiles. *Methods Ecol. Evol.* **12**: 2243–2255.
- Hazel, J. (2009) Evaluation of fast-acquisition GPS in stationary tests and fine-scale tracking of green turtles. *J. Exp. Mar. Biol. Ecol.* **374**: 58–68.
- Ibanez-Erquiaga, B., Pacheco, A. S., Rivadeneira, M. M. and Tejada, C. L. (2018) Biogeographical zonation of rocky intertidal communities along the coast of Peru (3.5–13.5 S Southeast Pacific). *PLoS One* **13**: e0208244.
- Irvine, L. M., Winsor, M. H., Follet, T. M., Mate, B. R. and Palacios, D. M. (2020) An at-sea assessment of Argos location accuracy for three species of large whales, and the effect of deep-diving behavior on location error. *Anim. Biotelemetry* **8**: 20.
- Kroodtsma, D. A., Mayorga, J., Hochberg, T., Miller, N. A., Boerder, K., Ferretti, F., et al. (2018). Tracking the global footprint of fisheries. *Science* **359**: 904–908.
- Kruger, L., Ramos, J. A., Xavier, J. C., Gremillet, D., Gonzalez-Solis, J., Petry, M.V., et al. (2018) Projected distributions of Southern Ocean albatrosses, petrels and fisheries as a consequence of climate change. *Ecography* **41**: 195–208.
- Lascelles, B. G., Taylor, P. R., Miller, M. G. R., Dias, M. P., Opper, S., Torres, L., et al. (2016) Applying global criteria to tracking data to define important areas for marine conservation. *Divers. Distrib.* **22**: 422–431.
- Louzao, M., Becares, J., Rodriguez, B., Hyrenbach, K. D., Ruiz, A. and Arcos, J. M. (2009) Combining vessel-based surveys and tracking data to identify key marine areas for seabirds. *Mar. Ecol. Prog. Ser.* **391**: 183–197.
- Majluf, P., Babcock, E. A., Riveros, J. C., Schreiber, M. A. and Alderete, W. (2002) Catch and bycatch of sea birds and marine mammals in the small-scale fishery of Punta San Juan, Peru. *Conserv. Biol.* **16**: 1333–1343.
- Merkel, B., Phillips, R. A., Descamps, S., Yoccoz, N. G., Moe, B. and Strom, H. (2016) A probabilistic algorithm to process geolocation data. *Mov. Ecol.* **4**: 26.
- Moreno, C. and Quiñones, J. (2022) Albatross and petrel interactions with an artisanal squid fishery in southern Peru during El Niño, 2015–2017. *Mar. Ornithol.* **50**: 49–56.
- Mott, R., and Clarke, R. H. (2018) Systematic review of geographic biases in the collection of at-sea distribution data for seabirds. *Emu* **118**: 235–246.
- Opper, S., Meirinho, A., Ramirez, I., Gardner, B., O'Connell, A.F., Miller, P., et al. (2012) Comparison of five modelling techniques to predict the spatial distribution and abundance of seabirds. *Biol. Conserv.* **156**: 94–104.
- Orben, R. A., Adams, J., Hester, M., Shaffer, S. A., Suryan, R. M., Deguchi, T., et al. (2021) Across borders: external factors and prior behaviour influence North Pacific Albatross associations with fishing vessels. *J. Appl. Ecol.* **58**: 1272–1283.
- Paulik, G. J. (1981) Anchovies, birds and fisherman in the Peru current. In M. H. Glantz and J. D. Thompson eds. *Resource management and environmental uncertainty: lessons from coastal upwelling fisheries.* New York, USA: John Wiley and Sons, 156–185.
- Peron, C. and Gremillet, D. (2013) Tracking through life stages: adult, immature, and juvenile autumn migration in a long-lived seabird. *PLoS One* **8**: e72713.
- Phillips, R. A., Silk, J. R. D., Croxall, J. P., Afanasyev, V. and Briggs, D. R. (2004) Accuracy of geolocation estimates for flying seabirds. *Mar. Ecol. Prog. Ser.* **266**: 265–272.
- Priddel, D., Carlile, N., Portelli, D., Kim, Y., O'Neill, L., and Bretagnolle, V. (2014) Pelagic distribution of Gould's Petrel (*Pterodroma leucoptera*): linking shipboard observations and onshore observations with remote-tracking data. *Emu* **114**: 360–370.
- Quiñones, J., Alegre, A., Romero, C., Manrique, M. and Vasquez, L. (2021) Fine-scale distribution, abundance, and foraging behavior of Salvin's, Buller's, and Chatham Albatrosses in the Northern Humboldt Upwelling System. *Pacific Sci.* **75**: 85–105.
- Quiñones, J., Calderon, J., Mayaute, L. and Bell, E. (2020) Black petrel (*Procellaria parkinsoni*) congregations at sea off Perú during the Austral summer. *Notornis* **67**: 573–576.
- R Core Team. (2021) *R: A language and environment for statistical computing.* Vienna, Austria: R Foundation for Statistical Computing.
- Raine, A. F., Gjerdrum, C., Pratte, I., Madeiros, J., Felis, J. and Adams, J. (2021) Marine distribution and foraging habitat highlight potential threats at sea for the Endangered Bermuda petrel *Pterodroma cahow*. *Endanger. Species Res.* **45**: 337–356.
- Rayner, M. J., Hauber, M. E., Steeves, T. E., Lawrence, H. A., Thompson, D.R., Sagar, P. M., et al. (2011) Contemporary and historical separation of transequatorial migration between genetically distinct seabird populations. *Nat. Commun.* **2**: 332.
- Rexer-Huber, K., Parker, G., Sagar, P. and Thompson, D. (2021) *Salvin's Albatross breeding dates and productivity: nest-camera analysis.* Wellington, NZ: Department of Conservation.

- Reynolds, R. W., Smith, T. M., Liu, C., Chelton, D. B., Casey, K.S. and Schlax, M. G. (2007) Daily high-resolution-blended analyses for sea surface temperature. *J. Clim.* **20**: 5473–5496.
- Richard, Y., Abraham, E. R. and Berkenbusch, K. (2020) *Assessment of the risk of commercial fisheries to New Zealand seabirds, 2006–07 to 2016–17*. Wellington, NZ: Fisheries New Zealand.
- Sagar, P., Charteris, M., Parker, G., Rexer-Huber, K. and Thompson, D. (2018) *Salvin's albatross: Bounty Islands population project. Ground component*. Wellington, NZ: Department of Conservation.
- Sansom, A., Wilson, L. J., Caldwell, R. W. G. and Bolton, M. (2018) Comparing marine distribution maps for seabirds during the breeding season derived from different survey and analysis methods. *PLoS One* **13**: e0201797.
- Schaefer, A. L., Lukacs, P. M. and Kissling, M. (2015) Testing factors influencing identification rates of similar species during abundance surveys. *Condor* **117**: 460–472.
- Shaffer, S. A., Tremblay, Y., Weimerskirch, H., Scott, D., Sagar, P. M., Moller, H., et al. (2006) Migratory shearwaters integrate oceanic resources across the Pacific Ocean in an endless summer. *Proc. Natl. Acad. Sci. U.S.A.* **103**: 12799–12802.
- Spear, L. B., Ainley, D. G. and Webb, S. W. (2003) Distribution, abundance, and behaviour of Buller's, Chatham Island, and Salvin's Albatrosses off Chile and Peru. *Ibis* **145**: 253–269.
- Spiegelhalter, D., Thomas, A., Best, N. and Lunn, D., (2007) OpenBUGS user manual. *Version*, 3(2).
- Taylor, G. A. (2000) *Action plan for seabird conservation in New Zealand. Part A: threatened seabirds*. Wellington, NZ: Department of Conservation.
- Thompson, D., Sagar, P., Briscoe, D., Parker, G., Rexer-Huber, K. and Charteris, M. (2020) *Salvin's albatross: Bounty Islands population project. Ground component*. Wellington, NZ: Department of Conservation.
- Thompson, D., Sagar, P., Torres, L. and Charteris, M. (2014) *Salvin's albatrosses at the Bounty Islands: at-sea distribution*. Wellington, NZ: Department of Conservation.
- Warwick-Evans, V., Atkinson, P. W., Walkington, I. and Green, J. A. (2017) Predicting the impacts of windfarms on seabirds: an individual-based model. *J. Appl. Ecol.* **55**: 503–515.
- Watanuki, Y., Suryan, R. M., Sasaki, H., Yamamoto, T., Hazen, E. L., Renner, M., et al. (2016) *Spatial ecology of marine top predators in the North Pacific: tools for integrating datasets and identifying high use areas*. Sydney, Australia: Pices.
- Yamamoto, T., Watanuki, Y., Hazen, E. L., Nishizawa, B., Sasaki, H. and Takashi, A. (2015) Statistical integration of tracking and vessel survey data to incorporate life history differences in habitat models. *Ecol. Appl.* **25**: 2349–2406.
- Zhang, J., Bell, E. A. and Roberts, J. O. (2020) Demographic assessment of black petrels (*Procellaria parkinsoni*) on Great Barrier Island (Aotea Island). Wellington, NZ: Fisheries New Zealand.
- Zydelis, R., Lewison, R. L., Shaffer, S. A., Moore, J. E., Boustany, A. M., Roberts, J. J., et al. (2011) Dynamic habitat models: using telemetry data to project fisheries bycatch. *Proc. R. Soc. Lond. B. Biol. Sci.* **278**: 3191–3200.