

Fragmented populations of the Vulnerable eastern hoolock gibbon *Hoolock leuconedys* in the Lower Dibang Valley district, Arunachal Pradesh, India

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Abstract We conducted a survey of the distribution and population status of the eastern hoolock gibbon *Hoolock leuconedys* in non-protected fragmented forest areas of the Lower Dibang Valley district of Arunachal Pradesh, India, during September 2010–June 2011. We estimated the minimum population density at seven study sites by direct observation and by counting the number of vocalizations heard from groups that were out of view. We used 1-ha belt transects (500 × 20 m) to estimate tree diversity in the study area. For two groups of gibbons we recorded their feeding patterns in forest fragments, using focal individual sampling, during May 2011. A total of 54 groups and three solitary individuals were recorded: 39 groups and three individuals through visual encounter and 15 groups through vocalization recognition. The mean group size recorded was $2.89 \pm SE 0.11$ (range 2–4). We counted a total of 289 trees, of 26 species and 17 families, in 18 belt transect surveys covering all non-protected fragments of the Lower Dibang Valley district. The gibbons' diet consisted of seasonal fruits and figs (29%), leaves (65%), seeds (2%) and flowers (0.5%). The major threats recorded at the study sites were habitat destruction and hunting. To protect the Vulnerable eastern hoolock gibbon, conservation measures will need to involve local communities.

Keywords Conservation status, distribution pattern, eastern hoolock gibbon, fragmented population, *Hoolock leuconedys*, India, Lower Dibang Valley

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Introduction

Hoolock gibbons (genus *Hoolock*) occur in forested areas of north-east India, Bangladesh, Myanmar and southern China (Harlan, 1934; Groves, 1972). Geographically the natural range of these apes extends from the Brahmaputra River east to the Salween River

(Lwin et al., 2011). The genus *Hoolock* comprises two distinct species, the eastern hoolock gibbon *Hoolock leuconedys* and the western hoolock gibbon *Hoolock hoolock*, which are separated based on differences in fur coloration (Mootnick & Groves, 2005; Geissmann, 2007). *H. leuconedys* was known to be distributed east of the Chindwin River to the Salween River in Myanmar and south-west Yunnan Province in China, at 1,067–1,219 m altitude (Groves, 1971), until it was observed in Arunachal Pradesh, India, between the Lohit River in the north and the mountains of Dafa Bum in the south (Das et al., 2006). More recently the species has been found in Sadiya Division, the easternmost part of Assam, south of the Dibang–Brahmaputra River system (Chetry & Chetry, 2010). The population, distribution, ecology and behaviour of the western hoolock gibbon have been studied in India and Bangladesh by Alfred & Sati (1990), Das et al. (2003), Mukherjee (1986) and Islam & Feeroz (1992). Preliminary studies have been conducted on the eastern hoolock gibbon to explore its distribution and population status in Arunachal Pradesh and Assam, India (Chetry et al., 2008, 2010; Chetry & Chetry, 2010). Population surveys of the eastern hoolock gibbon have also been conducted in China (Fan et al., 2011) and Myanmar (Brockelman et al., 2009; Walker et al., 2009). The eastern hoolock gibbon is categorized as Vulnerable on the IUCN Red List (Brockelman & Geissmann, 2008) and listed as a Schedule I species in the Indian Wildlife (Protection) Act, 1972. To inform future conservation and management planning we carried out a survey of the fragmented forest areas of the Lower Dibang Valley district, Arunachal Pradesh, India, to determine the current population and conservation status of *H. leuconedys* in this region.

Study area

The study was conducted at seven sites outside Mehao Wildlife Sanctuary, in Arunachal Pradesh (Fig. 1): six sites to the south-east and north-west of the Sanctuary (Injuno, Koronu, Delo, Hawaichapori, Horupahar, Iduli) and one site to the north-west (Chidu). These seven study sites lie within two administrative circles in the Lower Dibang Valley district: Koronu Circle and Roing Circle. The study sites are at 145–390 m altitude and are unclassified forests, not included in any management class and under the jurisdiction of the territorial forest officer. The main forest types recorded in the area are low hills and plains semi-evergreen

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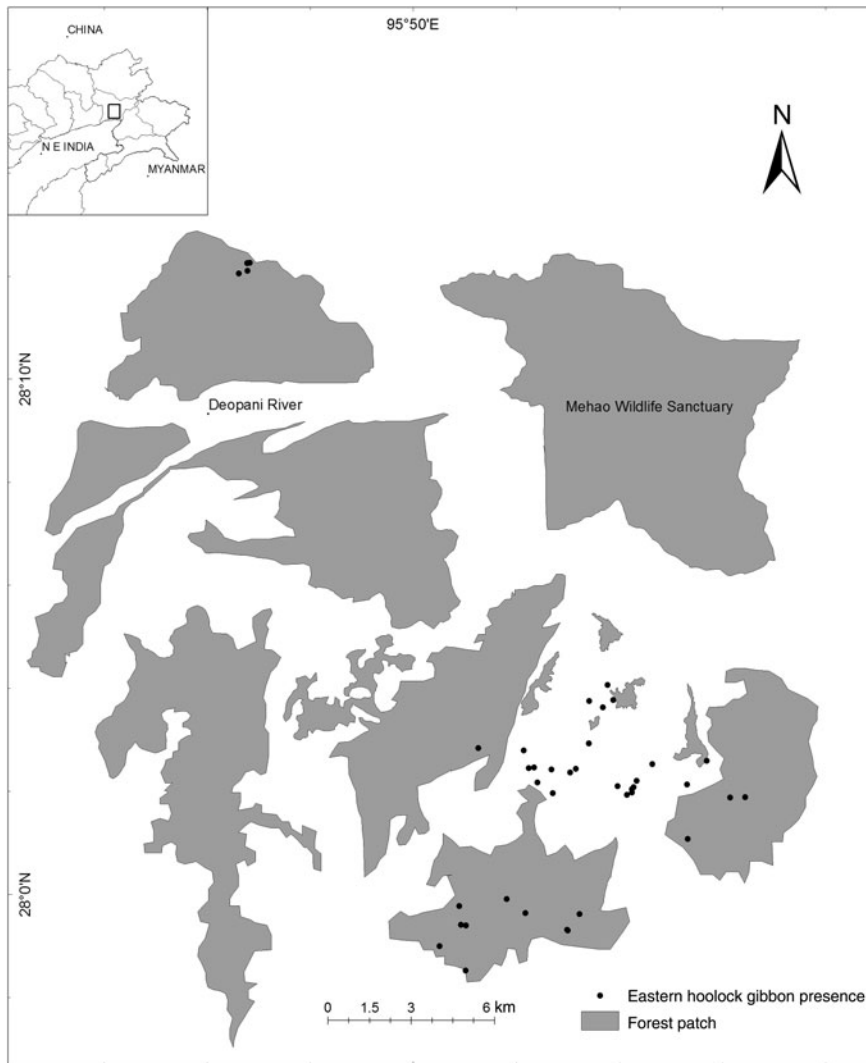


FIG. 1 The distribution of eastern hoolock gibbons *Hoolock leuconedys* and forest patches in the study area, in the Lower Dibang Valley district, Arunachal Pradesh, India.

forest, Assam alluvial plains semi-evergreen forest 2B/C1a and sub Himalayan light alluvial evergreen forest 2B/CI/ISI (Champion & Seth, 1968; Kaul & Haridasan, 1987). The study area is occupied by the Idu-Mishmi and Adi peoples, who generally depend on agriculture for their livelihood. They cultivate cash crops such as ginger *Zingiber officinale*, maize *Zea mays*, mustard *Brassica juncea* and rice *Oryza sativa*.

Methods

The survey was conducted from September 2010 to June 2011. We estimated the minimum population, based on visual encounters and auditory data, and recorded group size and age-class compositions (Kakati et al., 2009). We categorized individual gibbons as adult, subadult, juvenile or infant, based on body size and coat colour (Gupta et al., 2005; Table 1).

We covered a total of 109 km during 33 census walks across the seven study sites (Table 2). We sighted gibbons

directly from the transect and after following their calls (Kakati et al., 2009). We tried to locate all calling groups estimated to be < 500 m from the trail. For direct sightings we also recorded the local vegetation type and the degree of human disturbance of the habitat. We used stratified random sampling of transects to collect tree data within 18,500 × 20 m (1 ha; Sykes & Horrill, 1977; Kumar et al., 2006). We established the transects along existing forest trails and footpaths and recorded the local names and girth of all trees ≥ 30 cm in girth at breast height.

We estimated the minimum distance between gibbon groups, using the Distance Matrix tool in QGIS (QGIS, 2012). Using the geographical information system map we calculated the distance from each group to the nearest forest patch to examine the effect of forest degradation on group size. We counted the number of trees in the patches where gibbons were located, and assigned each patch to one of the following categories: single tree, single tree in bamboo patch, 2–5 trees, 5–7 trees, and > 7 trees. We compared the mean group size among the different categories. We analysed the

TABLE 1 Distinguishing characteristics of male and female eastern hoolock gibbons *Hoolock leuconedys* in different age categories (Gupta et al., 2005).

Age category	Sex	Distinguishing characteristics
Infants 1 day–2 years old	Male & female	20–35 cm length; carried by adult female; white to light grey (< 1 year), dark grey or black (1–2 years old); cannot determine sex in the field
Juveniles 2–4 years old	Male & female	35–44 cm length; black; distinct eyebrows; never carried by mother; sleep away from mother
Subadults 4–7 years old	Male	45–52 cm length; jet black; emerging scrotum
	Female	Broad greyish patches on a black coat
Adults > 7 years	Male	Distinct scrotum and jet black coat
	Female	Buff-coloured coat

change in forest cover, using satellite images of the study area from 1985 and 2010, with *ERDAS Imagine v. 9.2* (Intergraph, Madison, USA). We recorded the feeding and ranging patterns of two gibbon groups, A and B, collecting data on various aspects of food and feeding for each group for 7 consecutive days in May 2011. We recorded observations every 5 minutes, with focal individual sampling (Altmann, 1974). Each group comprised one adult male and one adult female with an infant. We recorded the time spent by each individual on each food plant, and the parts eaten, along with the time spent at different feeding sites (trees, climbers and crops). Plant parts were categorized as young leaves, mature leaves, fruits/figs, flowers and seeds. The groups were habituated to the presence of humans. We followed them from dawn to dusk each day. We focused on each adult in turn, to ensure representation of all members of the group. We used *SPSS v. 16.0* (SPSS, Chicago, USA) for statistical analysis, using Friedman and Mann–Whitney *U* tests to examine the differences in encounter rates and mean group sizes between the seven study sites.

Results

We estimated the total number of groups from the number of visual encounters and the vocal data. A total of 54 groups and three solitary individuals were recorded. The maximum number of groups was recorded at Iduli (17), followed by Delo (9), Horupahar (8), Koronu (7), Chidu (6), Injuno (5) and Hawaichapori (2). Of the 42 groups/individuals encountered directly, 38 groups were recorded in the Koronu Circle and four groups in the Roing Circle, at Chidu.

A total of 116 individuals in 39 groups and including 3 solitary individuals were recorded: 41 (35%) adult males,

38 (33%) adult females, 16 (14%) subadult males, 3 (3%) subadult females and 18 (15%) infants (Table 2).

The mean group size across the seven study sites was $2.89 \pm \text{SE } 0.11$ (range 2–4). For individual sites the highest mean group size was 3.0 (Koronu, Injuno, Hawaichapori and Chidu) and the lowest was 2.75 (Horupahar). There was a statistically significant difference in the group size and encounter rate across the seven sites (Friedman test, $\chi^2 = 10.57$, $P = 0.005$, $n = 7$). The ratio of immature (infant and subadult) to adult gibbons was the same at all sites. Group size also varied with habitat quality, specifically the number of standing trees (Fig. 2). Larger groups were found in forest patches (3.1) than in single trees (2.3). The distance of the group from the nearest forest patch also influenced the group size; smaller groups (2.5) were found at distances > 800 m (Fig. 3).

The estimated male : female ratio among adults was 1.07 : 1 across the seven study sites. The infant : female ratio was highest in the Injuno area (0.67 : 1; Table 2).

The recording of gibbons in Horupahar and Hawaichapori was significant because both these fragmented forest patches are surrounded by agricultural lands and human habitation, leaving gibbons more vulnerable at these sites. There was a significant correlation between the encounter rate and the size of the study area (Mann–Whitney *U*, $Z = 2.04$, $P < 0.05$, $n = 7$); furthermore, a higher encounter rate was observed when the mean distance between sites was lower (Fig. 4).

In the 18 transects we counted 289 trees, of 26 species (19 identified to species) and 17 families (Supplementary Table S1). The most common species were *Ficus* spp., followed by *Ailanthus grandis* and *Bischofia javanica*. The least frequently observed species was *Bombax ceiba*.

Satellite images of the study area for 1985–2010 show that the loss of forest cover was mainly as a result of the expansion of agricultural land, which increased by 87% (147 km²). In the study area a 48% decrease in forest cover was documented during 1985–2010 (Table 3).

Based on the total feeding time during a 7-day study period, the diet of the two focal groups of gibbons consisted of fruits and figs (29%), leaves (65%), seeds (2%) and flowers (0.5%). Gibbons showed a preference for young leaves (59%) over mature leaves (6%).

The main threats recorded at the study sites, through general observation and interaction with local people, were habitat destruction and hunting by local people. Habitat destruction was caused mainly by expansion of permanent agricultural practices, shifting cultivation, tea plantation, and construction of a national highway and permanent settlements. The eastern hoolock gibbon is hunted mainly as an alternative source of meat, and its bones, skin and fur are used for decoration and for making bags. Domestic dogs *Canis lupus familiaris* kept by local people to protect their property and livestock are potential

TABLE 2 Survey data from seven study sites in fragmented forest areas of Lower Dibang Valley district, Arunachal Pradesh, India (Fig. 1).

Variables	Chidu	Injuno	Horupahar	Hawaichapori	Delo	Koronu	Iduli
Altitude (msl)	390–430	211–283	195–215	198–216	209–251	212–259	151–179
Area of forest (km ²)	0.25	0.91	2.32	3.50	4.24	6.97	15.26
No. of survey days	4	4	5	2	5	5	8
Length of transect walked (km)	16.0	13.5	14.0	19.0	19.5	15.0	12.0
No. of groups encountered (G)	2	3	5	1	6	4	7
No. of solitaires encountered (S)	0	1	1	0	1	0	0
No. of groups heard/seen	2	0	3	0	1	1	4
Encounter rate (G+S per km)	0.13	0.22	0.35	0.05	0.30	0.26	0.58
Total no. of individuals	21	23	15	3	10	32	12
No. of adult individuals							
Male	8	8	5	1	4	11	4
Female	7	8	4	1	3	10	4
No. of subadult individuals							
Male	3	3	3	0	1	7	2
Female	0	0	0	0	0	3	0
No. of infants	3	4	2	1	2	4	2
Immature*/adult ratio	2.0	2.0	2.2	2.3	2.5	2.0	1.9
Infant/female ratio	0.50	0.67	0.50	1.00	0.43	0.40	0.40
Mean group size (No. of groups)	3.00 (4)	3.00 (3)	2.75 (8)	3.00 (1)	2.86 (7)	3.00 (5)	2.91 (11)
No. of groups heard only	2	2	0	1	2	2	6

*Infant + subadult

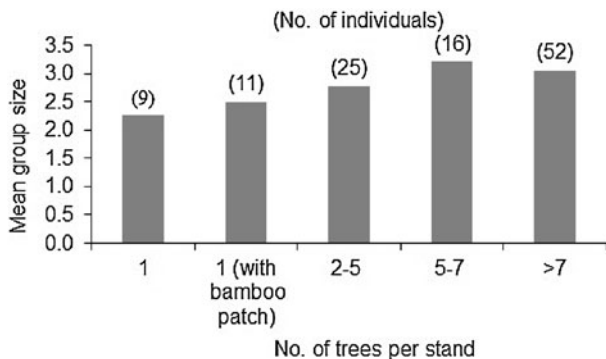


FIG. 2 Mean group size of eastern hoolock gibbons in stands of different numbers of trees in the study area, in the Lower Dibang Valley district, Arunachal Pradesh, India (Fig. 1). The numbers above the columns are the total numbers of individuals counted.

predators of gibbons, particularly of immature animals. Dogs are opportunistic predators, capturing gibbons when they descend to the ground to move from one patch to another in search of food and secure sleeping places.

Discussion

Gibbons are known to occur in the tropical evergreen, tropical wet evergreen, tropical semi-evergreen, tropical moist deciduous and subtropical hill forests of South-east Asia. The geographical range of the eastern hoolock gibbon in India is contiguous with its range in Myanmar through the Chukan pass in Changlang district (Das et al., 2006). The Chindwin River and the mountains on the India–

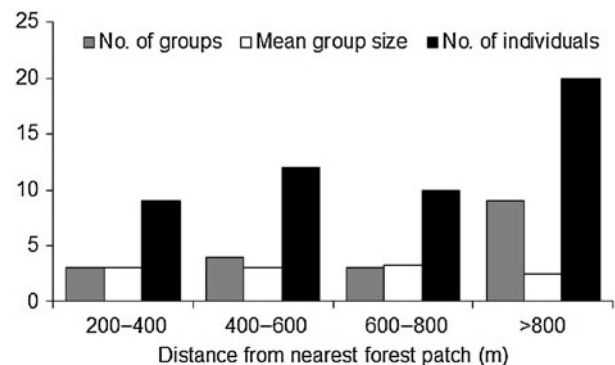


FIG. 3 Mean group size of eastern hoolock gibbons at various distances from the nearest forest patch in the study area, in the Lower Dibang Valley district, Arunachal Pradesh, India (Fig. 1).

Myanmar border are physical barriers to the distribution of the species. After the first distribution surveys of the eastern hoolock gibbon in India (Das et al., 2006; Chetry et al., 2008) a population survey was carried out in Mehao Wildlife Sanctuary, in the Lower Dibang Valley district of Arunachal Pradesh (Chetry et al., 2010). Gibbon occurrence in the area is dependent on the presence of the local Adi and Idu-Mishmi tribes; the Adi hunt gibbons, whereas the Idu-Mishmi do not. In this study we recorded gibbons in isolated trees and small patches of forest surrounded by agricultural land. The conversion of almost 50% of forest to agricultural land in the study area has led to canopy discontinuity, which prevents dispersal of the gibbons from one patch to another unless they descend to the ground. It also makes them more easily sighted by hunters and

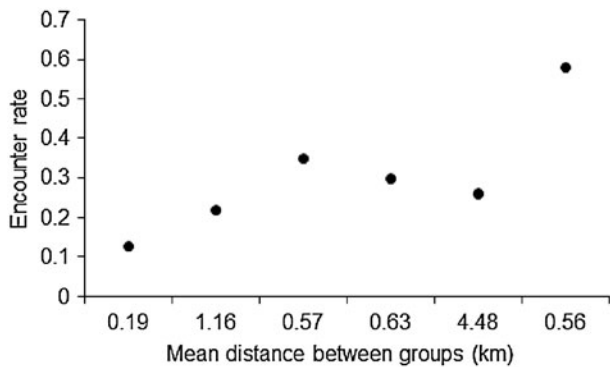


Fig. 4 The relationship between the encounter rate and the mean distance between groups of eastern hoolock gibbons in the study area, in the Lower Dibang Valley district, Arunachal Pradesh, India (Fig. 1).

potential predators such as domestic dogs and predatory birds such as the mountain hawk eagle *Nisaetus nipalensis*. Fragmentation also causes deterioration of the habitat and the elimination of suitable territories (Kakati et al., 2009) and reduces the availability of suitable mates and the genetic variability of the population (Jiang et al., 2006). The disappearance of gibbons from small forest fragments has been documented. Alfred & Sati (1990) recorded the disappearance of western hoolock gibbons from 168 forest patches (0.14–2.7 km²) in *jhum* (slash-and-burn agriculture) matrices in the Garo Hills of Meghalaya. This occurred mainly because the *jhum* cycles (Kushwaha et al., 1981) had shortened to < 10 years and gibbon dispersal corridors (secondary forests and old-growth bamboo) were no longer available.

Little information is available on the populations of the eastern hoolock gibbon in India. Other surveys have estimated 51 groups (168 individuals) in Namsai Forest Division, 10 groups (24 individuals) in Koronu Circle, 157 groups (> 88 individuals) in Mehao Wildlife Sanctuary, in Arunachal Pradesh, and 23 groups in Sadiya Forest Division, in Assam (Das et al., 2006; Chetry et al., 2008, 2010; Chetry & Chetry, 2010). Our study has added to the knowledge of the Indian population of the species and highlighted the conservation importance of the fragmented population in the non-protected forest areas of Arunachal Pradesh.

The mean group size varies between habitats and depends on the level of anthropogenic disturbance. The highest mean group size for eastern hoolock gibbon (3.37) was reported in the Namsai Forest Division, Lohit district (Das et al., 2006), followed by 3.14 and 2.4 in Lower Dibang Valley district in Arunachal Pradesh (Chetry et al., 2008, 2010). The smaller mean group size (2.89) in our study area relative to other parts of the species' range (e.g. 3.9 in China; Brockelman et al., 2009; Fan et al., 2011) may be a result of forest fragmentation and human disturbance. We found the ratio of infants to adult females to be similar in

all seven study sites (i.e. breeding rates were similar) but the absence of juveniles indicates that their survival may be affected by fragmentation and lack of food. Female gibbons in forest fragments probably suffer high lactation costs because of an inadequate diet (Kakati et al., 2009), both in quality and quantity. In such a situation dependent infants are more likely to die when they stop suckling. When resources are scarce females may not survive to breed again (Moir, 1994).

Chetry et al. (2008) reported forest loss and fragmentation in Lower Dibang Valley district as a result of extensive agricultural practices (e.g. tea, ginger, corn and mustard cultivation). We identified activities such as firewood collection, selective tree felling, and encroachment for permanent settlement as indirect threats that may affect the gibbon population. Almost 50% of the total forest loss has occurred during the last 25 years, restricting the gibbons to small fragments of forest. Single-tree groups are most at risk as they are smaller than groups in forest patches. Although hunting is not widespread in Idu-Mishmi tribal areas, even a low level of hunting can significantly affect the gibbon population (Fan & Jiang, 2007). Hunting pressure has apparently led to the local extinction of the gibbon in the Adi-dominated area. In a tribal state such as Arunachal Pradesh, where hunting is part of the local culture, it is difficult to implement wildlife laws and protect species listed under the Indian Wildlife (Protection) Act, 1972. In addition to hunting and poaching, attacks by domestic dogs are a threat (Panor, 2011). According to local people, domestic dogs sometimes attack gibbons, particularly immature animals that descend to the ground to move between forest patches.

Hoolock gibbons have been characterized as predominantly frugivorous (Chivers, 1974; Tilson, 1979; Alfred & Sati, 1994), with fruits constituting up to 70% of their diet (Chivers, 1984; Islam & Feeroz, 1992; Ahsan, 2001; Bartlett, 1999). However, our study indicated a more folivorous diet comprising 65% leaves and only 29% fruits and figs. Other studies have also recorded a significant reduction in the fruit content of the diet (Mukherjee, 1986; Kakati, 1997). Although it has been noted that folivory in hoolock gibbons increases during the wet (monsoon) season, when fruit tends to be less abundant (Gittins & Tilson, 1984; Alfred & Sati, 1994), our study was carried out in May, when fruits are normally more readily available than at other times. A study of the species' diet across seasons is needed to assess annual dietary variation.

Fan et al. (2011) recommended long-term population monitoring to determine whether hoolock gibbons can disperse between fragmented forest patches. For effective conservation management of gibbons and their habitat ongoing evaluation of their status is necessary, particularly in areas that hold a comparatively large proportion of the total population in India (Struhsaker et al., 1975; Wilson &

TABLE 3 Land use/land cover change matrix for the study area during 1985–2010. The numbers indicate the change in area (km²); e.g. 171.62 km² forested land was converted to cultivated land, and a total of 346.51 km² of forested land was converted to other land uses.

Sum of area (km ²) 1985	2010				
	Cultivated land	Water body	Forested land	Built-up area	Total
Cultivated land	137.20	2.25	28.90	1.41	169.76
Water body	8.18	25.33	6.07	0.00	39.59
Forested land	171.62	24.16	144.98	5.75	346.51
Built-up area	0.03	0.00	0.08	3.27	3.38
Total	317.04	51.74	180.02	10.42	559.23

Wilson, 1975). We recommend habitat improvement through reforestation, and construction of canopy bridges to connect remnant forest patches. The translocation of gibbons from the Delo area to forested areas in the Mehao Wildlife Sanctuary as a part of a conservation action plan that has been initiated by the Wildlife Trust of India, in collaboration with the Forest Department, Arunachal Pradesh, must focus on saving isolated populations or groups that are at risk of extirpation. However, the foremost requirement for better conservation of the species is increased awareness and involvement of the local communities in addressing the threats of forest cutting and agricultural expansion.

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