

The effects of a national highway on the Endangered golden-brown mouse lemur *Microcebus ravelobensis* in Ankarafantsika National Park, Madagascar

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Abstract Although roads are often assumed to be barriers to the dispersal of arboreal species, there has been little empirical testing of this assumption. If arboreal animals are unable to cross roads, population subdivision may occur, or resources may become inaccessible. We tested the hypothesis that Route Nationale 4 (RN4), a paved highway, was a barrier to movement and dispersal of the Endangered golden-brown mouse lemur *Microcebus ravelobensis* in Ankarafantsika National Park, in north-west Madagascar. During June–August 2015 we conducted a capture–mark–recapture study at three sites: two adjacent to RN4 and one within intact forest without a potential barrier. During 2,294 trap nights we captured 120 golden-brown mouse lemurs 1,032 times. In roadside habitats we captured significantly more males than females, whereas the opposite was the case in interior forest habitat. We detected eighteen crossings of highway transects by nine individuals; however, all potential dispersal events involved males. In roadside habitat, movement was significantly inhibited in both males and females. We present some of the first data on the effects of roads on movement patterns in arboreal Malagasy mammals, showing species- and sex-biased effects of roads as dispersal barriers. Our findings indicate that roads may not be complete barriers to dispersal in lemurs. We recommend that conservation managers and scientists examine explicitly the effects of roads and natural arboreal bridges in Madagascar in future studies.

Keywords Dispersal, lemur, Madagascar, *Microcebus ravelobensis*, movement, road ecology, sex-biased effects, vehicular mortality

Introduction

There is an extensive literature on the effects of roads on wildlife and environments (e.g. Forman, 2003; van der

Ree et al., 2015). However, many road ecology studies have focused on terrestrial species in temperate environments, especially in areas with highly developed road infrastructure, such as Europe and North America (Holderegger & Di Giulio, 2010). These studies have focused on large, paved freeways and finding practical solutions to the problems posed by roads (Smith et al., 2015). Arboreal animals face additional challenges, as they must descend to the ground to cross roads (Soanes & van der Ree, 2015). However, connectivity of habitat across roads, both natural and anthropogenic, can mitigate the negative effects of roads if arboreal animals traverse these connections (Van der Hoeven et al., 2010). There are relatively fewer data on how arboreal animals respond to roads in tropical forest regions, particularly for species found in rare tropical dry forests, which are amongst the world's most threatened ecosystems (Sagar & Singh, 2006). Thus, hypothesis-driven studies are needed in the tropics to determine the impact of roads, rather than an a priori assumption of roads as barriers or vice versa.

Road ecology in Madagascar is particularly important because 94% of lemur taxa are currently listed by IUCN as threatened with extinction (Schwitzer et al., 2014) and expanding road construction may be a major cause of lemur habitat loss (Rogers et al., 2010), although there has been little empirical testing of this assumption. The lack of data is surprising given that roads bisect or border some of the largest and best-known protected areas on the island, such as Ranomafana and Ankarafantsika National Parks.

We investigated whether individual golden-brown mouse lemurs *Microcebus ravelobensis* were able to cross roads, and how their movement was affected in road environments with various levels of canopy connectivity and uninterrupted forest habitat along a national highway in Ankarafantsika National Park. We predicted that (1) individual lemurs would be able to cross roads when canopy connectivity was present, (2) individual lemurs in uninterrupted forests would not be hindered in their ability to move, and (3) there would be no significant effect of sex on the ability of lemurs to cross roads, as both males and females of the species disperse. These predictions were evaluated to test the hypothesis that roads were a barrier to movement in *M. ravelobensis*.

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Methods

We conducted a capture–mark–recapture study during the dry season (May–August) of 2015 in the Ampijoroa Forest Reserve within Ankarafantsika National Park. There are various microhabitats within the Park (Rakotondravony & Radespiel, 2009), and Ampijoroa is characterized by infertile sandy soils with low water-holding capacities, gallery forest alongside Lake Ravelobe, and a high occurrence of non-native plantation trees, such as *Mangifera indica* and *Tectona grandis* (Rendigs et al., 2003; Crowley et al., 2012). Route Nationale 4 (RN4) is a 580 km two-lane paved highway from Antananarivo to Mahajanga, which runs directly through Ankarafantsika National Park in a north–south direction for c. 20 km (Fig. 1), dividing the Park into two sections, of 17,500 and 78,000 ha (Radespiel et al., 2008).

Mouse lemurs are small-bodied (30–70 g), nocturnal and solitary omnivores (Kappeler & Rasoloarison, 2003). Individuals of our study species, *M. ravelobensis*, establish mixed-sex sleeping sites, their diets contain a higher percentage of arthropods, and they have a slightly sex-biased system of female philopatry, but both sexes disperse (Radespiel et al., 2003, 2008; Weidt et al., 2004). Previous



FIG. 1 Major national highways (dashed lines) and protected areas (grey shading) in Madagascar. The inset shows the division of Ankarafantsika National Park by Route Nationale 4 (RN4) into areas of c. 17,500 and 78,000 ha.

studies have suggested *M. ravelobensis* populations may be affected by roads, but because the road construction was relatively recent, no explicit genetic barrier was found (Radespiel et al., 2008). Mouse lemurs are excellent model organisms to distinguish between the direct and secondary effects of roads because they are under less direct hunting pressure than larger-bodied animals (Garcia & Goodman, 2003).

We established six parallel trap lines of 475 m length at three locations in Ampijoroa (Fig. 1): two on either side of RN4 where there was canopy connectivity across the highway (Road 1), two on either side of RN4 at a site with no canopy connectivity (Road 2), and two within the interior forest, c. 1 km from the road or forest edge, as a control (Forest Control). We sampled trap lines on a 6-day randomized schedule; no trap line was sampled on 2 consecutive nights and monthly trapping intensity was equal at all sites.

We set live animal traps (Sherman, Tallahassee, USA) at 25 m intervals along each trap line, giving a total of 20 trapping sites per trap line and 40 traps at each location. We established the parallel interior transects at similar distances to those on the roads, c. 25 m apart. Upon capture, individuals were categorized by sex and age-class and were marked subcutaneously with a Trovan microchip. For a more detailed description of trapping procedures see Burke & Lehman (2014). Our study was approved by the University of Toronto Office of Research Ethics (Protocol #20011050) and complied with all legal requirements of Madagascar.

A crossing event was defined as when an individual was captured on one trap line and then recaptured on the parallel trap line. We used *t*-tests, χ^2 tests, a one-way ANOVA, and Tukey's post-hoc analysis with $\alpha = 0.05$ to test for differences in movement and capture rates between locations and sex categories. Movement was quantified by rounding distances between adjacent non-diagonal traps to 25 m. All tests were performed in R v. 3.2.2 (R Development Core Team, 2013).

Results

We captured 120 *M. ravelobensis* individuals in 1,032 capture events at the three locations over 2,245 trap nights (2,520 traps set, with 82 rodent captures and 144 trap malfunctions). There were no significant differences between trapping locations in terms of rates of rodent captures ($t = 0.34$, $df = 60$, $P = 0.71$) or trap malfunctions ($t = 1.99$, $df = 60$, $P = 0.14$). There were significantly more males caught at Road 1 and Road 2 than at Forest Control; males were also caught at a higher rate than females at Road 1 and Road 2, whereas at Forest Control there were no significant differences in the number of individuals caught by sex (Table 1).

We found no significant differences in mean distance travelled by males and females at any of the three trapping locations (Fig. 1, Table 2). There was a significant difference in total movement between the three locations (Table 2). Tukey's post hoc analysis indicated there was a significant difference between Forest Control and Road 1 ($Q = 5.68$, $df = 2,117$, $P < 0.001$) and Road 2 ($Q = 7.78$, $df = 2,117$, $P < 0.01$), but no significant difference between Roads 1 and 2 ($Q = 2.21$, $df = 2,117$, $P > 0.05$).

We detected road crossing on 18 occasions, by 9 males (Table 3). Of these 18 captures, 14 (seven individuals) were at Road 1. At Road 2 we captured two individuals four times. Individuals ($n = 5$) at both locations were recaptured on their original trap line after being caught on the parallel line. At Forest Control we recaptured *M. ravelobensis* 157 times on a different trap line than that on which they were previously caught, corresponding to 77% ($n = 27$ of 35) of individuals.

Discussion

Our study is the first to directly test the permeability of roads to mammal movement in Madagascar, and our findings indicate that RN4 affects the movement of *M. ravelobensis* in Ankarafantsika National Park. Our first two predictions were validated: although the lemurs were able to cross RN4, they did so at a lower rate than if there had been undisturbed canopy. Thus, it is important for conservation managers to protect existing crossings and encourage more canopy growth, as our results suggest that crossing movement of *M. ravelobensis* would be further reduced if existing arboreal bridges were lost. However, Madagascar is prone to extreme climatic events, such as cyclones, droughts and wildfires (Ingram & Dawson, 2005), and in

TABLE 1 Number of individuals captured (individuals identified and given a unique ID) and total number of captures (including recaptures) of golden-brown mouse lemurs *Microcebus ravelobensis* at the Road 1 (connection across RN4), Road 2 (no connection across RN4) and Forest Control (interior forest with no barrier) sites in the Ampijoroa Forest Reserve within Ankarafantsika National Park, Madagascar (Fig. 1), with χ^2 and P (* denotes significance at $P < 0.05$).

Location	Males	Females	χ^2	P	Total
Individuals captured					
Road 1	29	15	4.45	0.03*	44
Road 2	28	13	6.25	0.01*	41
Forest Control	17	18	0.11	0.74	35
<i>Total</i>	74	46			120
Total captures					
Road 1	177	101	20.78	0.01*	278
Road 2	227	117	45.30	< 0.01*	344
Forest Control	183	227	2.36	0.31	410
<i>Total</i>	587	445			1,032

TABLE 2 Distance moved (\pm SD) by male and female individuals of *M. ravelobensis* at the Road 1 (connection across RN4), Road 2 (no connection across RN4) and Forest Control (interior forest with no barrier) sites in the Ampijoroa Forest Reserve (Fig. 1), with results of t -tests and ANOVA (* denotes significance $P < 0.05$).

Location	Distance moved \pm SD (m)		t	df	P
	Males	Females			
Road 1	123 \pm 144	112 \pm 126	0.26	42	0.79
Road 2	208 \pm 166	175 \pm 203	0.54	39	0.59
Forest Control	299 \pm 220	428 \pm 290	1.48	33	0.15
<i>Total</i>	183 \pm 182	262 \pm 259	16.25	2,117	< 0.01*

2000 Cyclone Hudah resulted in the destruction of 51% of trees and an 83% reduction in mean tree crown volume in the Masoala Peninsula (Birkinshaw & Randrianjanahary, 2007). Cyclones of a similar nature and intensity occur in Ankarafantsika National Park (Aymoz et al., 2013). Additionally, if construction occurs along RN4, anthropogenic damage to these natural arboreal bridges may also occur (Weller, 2015). Thus, it would also be prudent for managers to investigate the efficacy of permanent anthropogenic crossing structures, such as bridges or culverts.

Our third prediction was not supported: only male *M. ravelobensis* were observed crossing RN4 and there were significantly more males alongside RN4 than within the interior forest. Although we captured significantly more males than females at Roads 1 and 2, we captured 15 and 13 individual females, respectively, over 100 times at each location. Additionally, at Forest Control we observed an equivalent number of crossings by males and females (13 and 14, respectively), and thus we consider our results to be robust. A sex-biased effect of roads has also been observed in North American freshwater turtles (Order Testudines), with female turtles more likely to be killed by vehicles than males while crossing roads to find nesting sites, leading to skewed sex ratios (Steen et al., 2006).

TABLE 3 The number of individuals of *M. ravelobensis* captured on both parallel trap lines (individuals crossing), and the total number of captures in which the previous capture was on the parallel trap line (crossing events), at the Road 1 (connection across RN4), Road 2 (no connection across RN4) and Forest Control (interior forest with no barrier) sites in the Ampijoroa Forest Reserve (Fig. 1).

Location	No. of individuals crossing			No. of crossing events		
	Males	Females	Total	Males	Females	Total
Road 1	7	0	7	14	0	14
Road 2	2	0	2	4	0	4
Forest Control	14	13	27	70	87	157
<i>Total</i>	23	13	36	88	87	175

However, our results indicate that male *M. ravelobensis* are more likely to face mortality, as the crossing sex. Given that male and female *M. ravelobensis* have home ranges of equivalent size and both sexes disperse, this effect must be attributable to other factors, such as competition between males (Weidt et al., 2004) or home range pile-up, in which the lemurs are not affected by roads as direct barriers to dispersal, but through many individual or group home ranges abutting roads (Riley et al., 2006). This phenomenon can have significant genetic and social consequences in gregarious species such as mouse lemurs.

Our results on the roadside movement of *M. ravelobensis* are some of the first such data available for mammals in Madagascar, and contribute valuable data on the effects of roads on the understudied guild of arboreal mammals, for which research is especially needed in tropical environments such as Madagascar (Lehman et al., 2016). Managers of protected areas such as Ankarafantsika that are intersected or abutted by roads should investigate the efficacy of mitigation measures such as bridges or culverts (Smith et al., 2015) in aiding the movement of wildlife. However, anthropogenic bridges are potentially costly for developing nations, and have shown mixed results in aiding movement across barriers in Madagascar specifically (Mass et al., 2011). Our results suggest that natural bridges can aid movement, but further study comparing the efficacy of natural and anthropogenic bridges along RN4 and other roads in Madagascar is needed to clarify which is more effective. Given the urgent conservation needs and the rapid development of roads within Madagascar, more hypothesis-driven studies are needed on the effects of roads, to inform conservation initiatives.

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Author contributions

MSR conducted all data analysis and drafted the article. MSR and AR collected the data, assisted by SML. AR and

SML assisted in designing the study and preparing the article.

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Biographical sketches

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